



Serpent River Watershed Monitoring Program 2017 Annual Water Quality Report

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2017 SRWMP Annual Water Quality Report

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1 Introduction

As part of the closure and decommissioning process, Rio Algom Limited (RAL) and Denison Mines Inc. (DMI) developed a focused and integrated performance monitoring network for legacy sites within the Serpent River Watershed (SRW). The comprehensive monitoring and management strategy clearly defined and delineated the purpose for all monitoring activities through three integrated programs; the Tailings Management Area (TMA), the Operational Monitoring Program (TOMP), the Source Area Monitoring Program (SAMP), and the Serpent River Watershed Monitoring Program (SRWMP) (Minnow Environmental Inc. (Minnow), 2016). An integrated assessment of the results from all of these programs is prepared every five years in a *State of the Environment Report* (SOE) in compliance with license requirements and in accordance to Canadian Standards Association (CSA) N288.4-10 (2010). The regulatory review draft of the most recent SOE covering data collection and monitoring for the period of January 1, 2010 – December 31, 2014 was issued to the Canadian Nuclear Safety Commission (CNSC) as well as other members of the Joint Regulatory Review Group (JRG) on March 10, 2016 with response to regulatory comments issued September 30, 2016. The final SOE report with regulatory approval was distributed in November 2017.

The SRWMP was initiated in 1999 as a joint initiative of RAL and DMI with the objectives of evaluating the effectiveness of mine decommissioning plans and assessing long-term environmental quality trends in the watershed (Beak International Incorporated (Beak), 1999). Evolution of the program, key outcomes, program modification decisions, and associated references are summarized in Appendix I. In 2017, the SRWMP followed the 2016 approved program modifications recommendations described in the document submission entitled *Cycle 4 Study Design for the SRWMP, SAMP and TOMP* (Minnow, 2016).

The SRWMP Annual Water Quality Report for 2017 provides water quality data from shared RAL and DMI watershed monitoring locations from January 1, 2017 thru to December 31, 2017. This report should be read in conjunction with the Annual Operating, Care and Maintenance (OCM) Reports, prepared independently by each company, that incorporate upstream SAMP and TOMP data, and discuss operational activities of each company (RAL, 2018; DMI, 2018). The objective of the SRWMP annual data review is to identify anomalous data and provide visual evaluation of short-term data trends at key locations. Step changes and anomalies are identified in this report by reviewing and compiling the last five years of annual average data for all SRWMP monitoring locations, and visually reviewing the information for any noticeable changes. Significant changes and unusual results are investigated in accordance with the *Water Quality Assessment and Response Plan*, which is found in Appendix A of the most recent SOE Report (Minnow, 2017).

The SRWMP Annual Water Quality Report for 2017 also provides a summary of the quality management program and water quality results for the period January 1, 2017 through December 31, 2017.

As part of the 2015 SOE review, CNSC has instructed RAL and DMI to include annual reporting of a representative radiation dose to the public associated with their closed uranium mine sites in the Serpent River Watershed. Details on this topic are included in Section 4.4 of this report.

2 METHODOLOGY

2.1 2017 Program Requirements

The 2017 SRWMP followed program requirements (sampling locations, frequencies, parameters, and analytical protocols) as recommended and approved in the *Cycle 4 Study Design for the*

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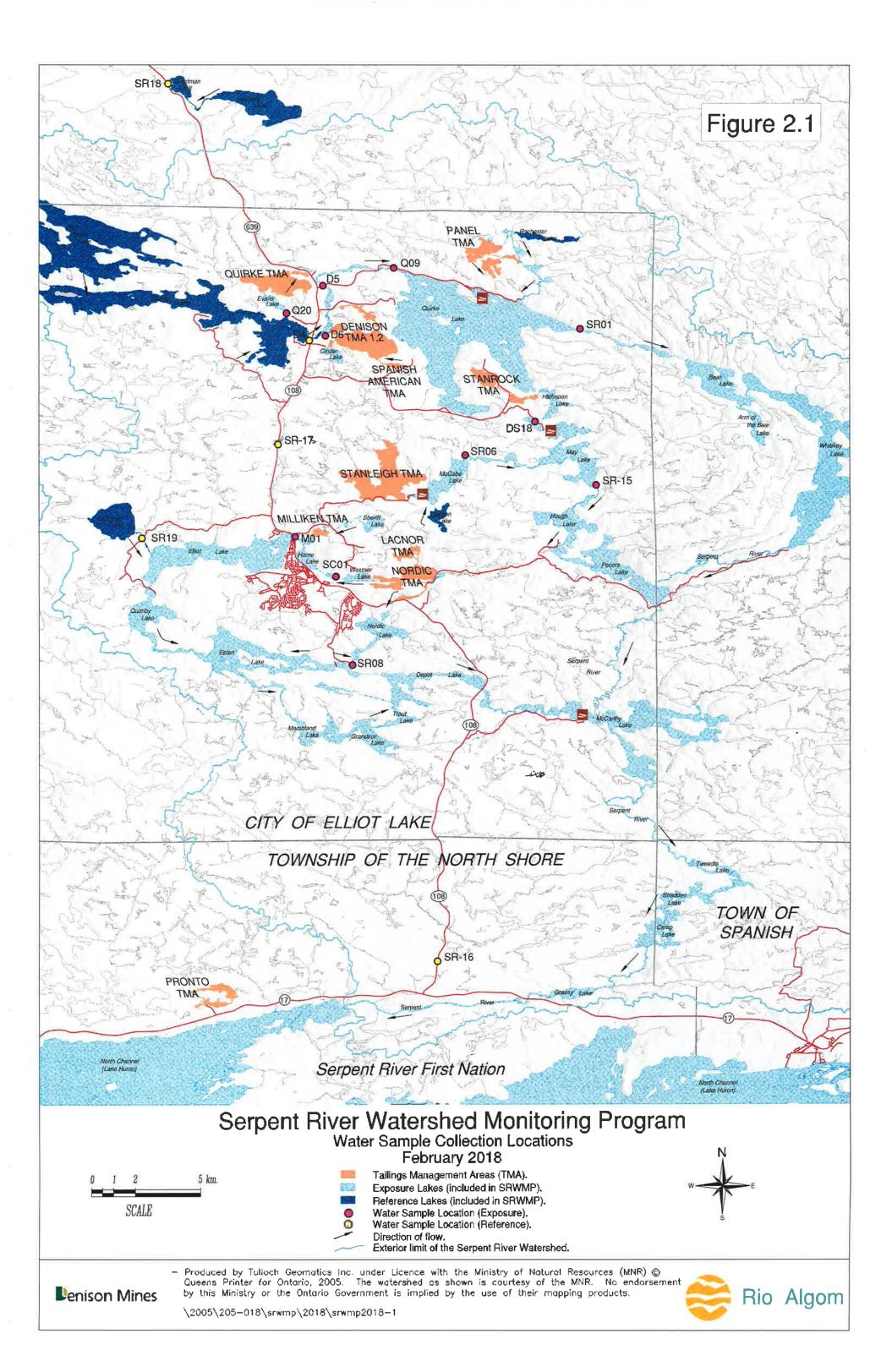
SRWMP, *SAMP* and *TOMP* (Minnow, 2016) as well as Environmental Compliance Approvals for Nordic (Ministry of Environment, 2009). Table 2.1 provides a brief description of each location, the frequency and parameters monitored as well as non-SRWMP regulatory drivers, and Figure 2.1 provides a map of the stations included in the water quality monitoring program.

Table 2.1 2017 SRWMP Water Quality Monitoring Requirements

Sampling Station	Location / Description	Sample Type	Purpose	Flow (L/s)	Field pH	Sulphate (mg/L)	Radium-226 (Bq/L total)	Urnaium (mg/L)	Barium (mg/L)	Cobalt (mg/L)	Iron (mg/L)	Manganese (mg/L)	Hardness² (mg/L)
SR-16 ²	Fox Creek at Highway 108	Wetland/stream reference	SAMP		4	4	4	4	4	4	4	4	4
SR-17 ²	Unnamed Creek Drain Lake 3 @ Hwy 108	Wetland/stream reference	SAMP		4	4	4	4	4	4	4	4	4
SR-18	Outlet of Jim Christ Lake	Lake reference	SRWMP		2	2	2	2	2	2	2	2	2
SR-19	Inlet to Elliot Lake	Lake reference	SRWMP		4	4	4	4	4	4	4	4	2
SR-08	Nordic Lk Outlet	far field	SRWMP & MOE		4	4	4	4	4	4	4	4	4
SR-15	May Lake Outlet	far field	SRWMP/ Internal	2	2	2	2	2	2	2	2	2	2
M-01 ¹	Sherriff Ck @ Hwy 108	near field	SRWMP		4	4	4	4	4	4	4	4	4
Q-09	Serpent River Below Q Effluent	near field	SRWMP	4	4	4	4	4	4	4	4	4	4
Q-20	Evans Lk Outlet to Dunlop Lk	near field	SRWMP	1	1	1	1	1	1	1	1	1	1
SC-01	Westner Lk Outlet	near field	SRWMP&MOE		1	1	1	1	1	1	1	1	1
SR-06	McCabe Lk Outlet	near field	SRWMP	2	2	2	2	2	2	2	2	2	2
FBR5	Field Blank Rio	QA/QC	SRWMP		2	2	2	2	2	2	2	2	2
BSR5	Blind Sample Rio	QA/QC	SRWMP		2	2	2	2	2	2	2	2	2
Rio Alg	om total excluding field blanks & blind sa	mples		9	32	32	32	32	32	32	32	32	30
D-4	Dunlop Lk Outlet	Lake reference	SRWMP		2	2	2	2	2	2	2	2	2
D-5	Serpent R. between Q and D	near field	SRWMP	4	4	4	4	4	4	4	4	4	4
D-6 ¹	Cinder Lk Outlet	near field	SRWMP	4	4	4	4	4	4	4	4	4	4
DS-18	Halfmoon Lk Outlet	near field	SRWMP&MOE	4	4	4	4	4	4	4	4	4	4
SR-01	Quirke Lk Outlet	far field	SRWMP		1	1	1	1	1	1	1	1	1
FBD2	Field Blank Denison	QA/QC	SRWMP		2	2	2	2	2	2	2	2	2
BSD2	Blind Sample Denison	QA/QC	SRWMP		2	2	2	2	2	2	2	2	2
Denisor	total excluding field blanks & blind sam	ples		12	15	15	15	15	15	15	15	15	15
Total QA	/QC samples			0	8	8	8	8	8	8	8	8	8
TOTAL	SAMPLES			21	47	47	47	47	47	47	47	47	45
QA/QC	Fraction of Total			0%	17%	17%	17%	17%	17%	17%	17%	17%	18%
Denison	Fraction of Total			57%	32%	32%	32%	32%	32%	32%	32%	32%	33%

Notes

- Field QA-QC designated stations.
- 2. SR-16 and SR-17 are part of SAMP program Cycle 4 but are historically SRWMP locations.



2.2 2017 Program Conformance

All Cycle 4 approved sampling, field measurement, and analytical requirements were met during the 2017 reporting period. Hardness continues to be monitored as an ancillary parameter at all SRWMP stations. According to the most recent Ambient Water Quality Guidelines from the British Columbia Ministry of Envrionment (BCMOE), manganese and sulphate are hardness dependent (BCMOE, 2006, 2013). Semi-annual response monitoring also continued downstream at the outlet of May Lake (SR-15) effective January 1, 2016. Response monitoring is discussed in further detail in Section 4.1.

2.3 Field Measurements

Field measurement requirements and protocols for the 2017 SRWMP are presented in detail in the *Cycle 4 Study Design for the SRWMP, SAMP and TOMP* (Minnow, 2016). Field Staff have been thoroughly trained and have reviewed procedures associated with the proper calibration and use of field equipment for the measurement of field parameters. The models and accuracy for equipment used in measuring SRWMP field parameters are provided in Table 2.3.

Table 2.3 SRWMP Field Equipment Models and Accuracy

Parameter	Meter	Accuracy	Unit
pН	YSI Pro 10	+/- 0.02	pH units
flow	Global Flow Probe	0.1	feet per second

2.4 Data Quality Objectives

Field and laboratory data quality objectives (DQOs) for the 2017 SRWMP are presented in detail in the *Cycle 4 Study Design for the SRWMP, SAMP and TOMP* (Minnow, 2016). Table 2.4.a. provides a summary of field DQOs and Table 2.4.b. provides a summary of laboratory methods, detection limits and DQOs.

Data quality assessment results are covered in Section 3 of this report.

Table 2.4.a. 2017 SRWMP Field Data Quality Objectives

		Assessme		Data Quality Objectives ²					
Parameter	Units	PWQO	Background	Detection Limit	Minimum ³	Field Blank Criteria	Field Precision		
		BCMOE		Lillit	Detectable Difference	Onteria			
Field Parameters ³									
Flow	L/s	-	-	method	method	-	30%		
pН				0.1	0.01 or 0.02	-	10%		
Lake Stations		6.5	-						
Wetland/Streams		-	5.2						
Laboratory Paramet	ers								
Barium	mg/L	1.0	-	0.005	-	0.01	20%		
Cobalt	mg/L	0.0025	-	0.0005	-	0.001	20%		
Iron	mg/L	-	-		-				
Lake Stations		-	0.49	0.02	-	0.04	20%		
Wetland/Streams		-	1.69	0.02	-	0.04	20%		
Manganese 4	mg/L	0.8	-	0.002	-	0.004	20%		
Radium (total)	Bq/L	1.0	-	0.005	-	0.01	20%		
Sulphate 4	mg/L	128-429	-	0.1	-	0.2	20%		
Uranium	mg/L	0.015	-	0.0005	-	0.001	20%		
Hardness	mg/L	-	-	0.5	-	1.0	20%		

Notes:

^{1.} Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

^{2.} Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

^{3.} Minimum detectable difference as identified in instrument manual

^{4.} Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

Table 2.4.b. 2017 SRWMP Laboratory Methods and Data Quality Objectives

		Assessme	nt Criteria ¹		Laboratory Data Quality Objectives ²						
Parameter	Units	PWQO BCMOE	Background	Method	Detection Limit	Laboratory Blank	Precision	Spikes	Accuracy (CRM)		
Barium	mg/L	1.0	-	ICP-MS	0.005	0.01	10%	20%	20%		
Cobalt	mg/L	0.0025	-	ICP-MS	0.0005	0.001	10%	20%	20%		
Iron	mg/L	-		ICP-OES							
Lake Stations			0.49		0.02	0.04	10%	20%	20%		
Wetland/Streams			1.69		0.02	0.04	10%	20%	20%		
Manganese 3	mg/L	0.8	-	ICP-MS	0.002	0.004	10%	20%	20%		
Radium (total)	Bq/L	1.0	-	Alpha Spectroscopy	0.005	0.01	20%	20%	-		
Sulphate ³	mg/L	128-429	-	lon Chromatography	0.1	0.2	10%	20%	20%		
Uranium	mg/L	0.015	-	ICP-MS	0.0005	0.001	10%	20%	20%		
Hardness	mg/L	-	-	ICP-OES	0.5	0.1	10%	-	_		

Notes:

- 1. Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)
- 2. Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)
- 3. Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

2.5 Changes in Analytical Methods

There were no changes in analytical methodology in 2017.

2.6 Reporting of Method Detection Limits

Program method detection limits (MDLs) are presented in Tables 2.4.a. and 2.4.b. The target MDL for radium-226 (0.005 Bq/l) was not met on all samples analysed in 2017 due to decreased sample throughput of the analytical laboratory. There was no change in method during this period; however, the laboratory was only able to claim an MDL of 0.007 Bq/L as a result of a higher standard deviation on the previous year's QC samples. Analytical performance will be reviewed with the laboratory in 2018 in efforts to re-establish the target MDL of 0.005 Bq/L.

2.7 Data Screening and Assessment Conventions

Data validation was conducted on SRWMP water quality data throughout the year. The assessment-screening process flags all data points outside a rolling minimum 12 value mean \pm 3 standard deviations.

Flagged data and short-term response plans for the SRWMP are reported quarterly to the regulatory agencies as part of the water quality report. Data validation of "flagged data" for the year 2017 can be found in Appendix I.

Annual water quality reporting is designed to be concise and focused on the presentation of data in a standardized format with limited interpretation, as per Section 14.2 of the Implementation Document (Beak, 1999c). Data validation ensures prompt response to upset conditions or unusual results, as documented in *Data Validation Procedures* in conjunction with *Water Quality Assessment and Response Plan*, which is included in Appendix B of the SOE (Minnow, 2017). Assessment criteria as outlined in Table 2.4.a. and 2.4.b. of this report, are standardized to approved benchmarks selected, rationalized and presented in Tables 4.3 and 4.5 of the *Cycle 4 Study Design for the SRWMP, SAMP and TOMP* (Minnow, 2016).

Approved program modifications implemented in January of 2015 focused water quality monitoring on lakes located immediately downstream of the decommissioned TMAs. A more in-depth and detailed statistical evaluation of water quality trends is included in the SOE every five years (Minnow 2009, 2011, 2017).

The most recent five-year annual concentrations of mine indicator parameters at key downstream locations are reviewed in this report in Figures 3.1.a to 3.1.c. Detailed statistical analysis of the SRW water quality reported between January 1, 2010 – December 31, 2014 is included in Appendix E of the final SOE Report issued to the CNSC and other members of the JRG in November, 2017.

3 RESULTS

3.1 Data Quality Results and Assessment

Detailed laboratory quality assurance and quality control (QA/QC) results are provided in Appendix III, and detailed field QA/QC results are provided in Appendix IV. Field quality control results are summarized in Table 3.1. Data quality assessments for each type of data quality objective are provided in the following sections.

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3.1.1 Laboratory Quality Assurance and Quality Control

In 2017, all analytical requirements for the SRWMP were contracted to laboratories with Canadian Association for Laboratory Accreditation Inc. (CALA) accreditations. Detailed laboratory QA/QC results are provided in Appendix III. The 10% objective for QA/QC was met by both labs. SGS performed 7769 analyses with 6685 QC checks, which represents 86% QC for sample analysis (Appendix III). The Elliot Lake Research Field Station (ELRFS) analyzed 101 batches totaling 1197 radium samples with each batch incorporating blank, certified reference material (CRM), duplicate, and spiked samples providing greater than 25% quality control checks (Appendix III).

3.1.2 Quality Assurance and Quality Control Resolution of Key Issues

There were no major issues with laboratory analysis requiring resolution in 2017. However, the radium target MDL of 0.005 Bq/L was not achieved by ELRFS, but the MDL still remained below the laboratory Data Quality Objective (DQO) of 0.01 Bq/L at <0.007 Bq/L (Appendix II).

3.1.3 Analytical Blank Performance

Laboratory quality control results confirm that blank data quality objectives were met for all parameters in all samples (Appendix III).

3.1.4 Analytical Duplicate Performance

Laboratory quality control results confirm that duplicate data quality objectives of 20% for radium and 10% for all other remaining parameters were achieved in all samples (Appendix III).

All analytical duplicate results were within the data quality objectives (Appendix III).

3.1.5 Analytical Laboratory Spike Performance

Laboratory quality control results confirm that the spike data quality objective of 20% was achieved for all parameters in all samples (Appendix III). It should be noted that the required SRWMP reporting detection limit for barium is 0.005 mg/L while the reporting limit for iron is 0.02 mg/L. However, the SGS reporting detection limit for barium is 0.0001 mg/L with a spike concentration of 0.005 mg/L and the reporting limit for iron is 0.007 mg/L with a spike concentration of 0.01 mg/L (Appendix III). As a result, the spike concentrations are equivalent to the reporting detection limits so most spikes result in non-detects. This results in the statistics indicating that the spike data quality objectives were not met; however, when the lower detection limit is applied, the spike data quality objective is achieved, indicating that the method is sound and reliable.

3.1.6 Analytical Certified Reference Material Performance

Laboratory quality control results confirm that the CRM data quality objective of 20% accuracy was achieved for all parameters in all samples in 2017 (Appendix III).

3.1.7 Field Blank Performance

Field Blank quality control results confirm that SRWMP field blank data quality objectives were achieved in 2017 (Appendix IV).

3.1.8 Field Precision Performance

Field precision quality control results confirm that SRWMP field precision data quality objectives were achieved in 2017 (Appendix IV).

Table 3.1 2017 SRWMP Field Quality Control Results Summary

QA/QC	рН	SO4	Ra(T)	U	Ва	Со	Fe	Mn
		(mg/L)	(Bq/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)
MDL ¹	-	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002
Field Blank Statistics								
Count	4	4	4	4	4	4	4	2
Average	5.3	<0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	0.002
Max	5.5	<0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	0.002
Min	5.2	<0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	0.002
Field Blank Exceedances								
Criteria ¹	-	0.2	0.01	0.001	0.01	0.001	0.04	0.004
Exceedance	0	0	0	0	0	0	0	0
Field Precision Statistics								
Count	4	4	4	4	4	4	4	2
Average	0.0%	3.3%	2.9%	2.5%	2.0%	0.0%	5.8%	5.1%
Max	0.0%	7.4%	11.8%	5.1%	8.0%	0.0%	8.3%	12.9%
Min	0.0%	0.0%	0.0%	0.0%	0.0%	0.0%	2.5%	1.1%
Field Precision Exceedances								
Criteria ¹	20%	20%	20%	20%	20%	20%	20%	20%
Exceedance	0	0	0	0	0	0	0	0

Notes:

¹ Data Quality Objectives taken from Table 5.2 af the Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016) Bold indicates an exceedance in the Data Quality Objectives (DQO's)

3.2 Location Summary

Annual average concentrations of SRWMP parameters for 2017 in comparison to the *Cycle 4 Study Design* (Minnow, 2016) receiving environment assessment criteria are provided in Table 3.2. Annual detailed results and five-year summaries of annual average concentrations in comparison to assessment criteria are provided in Appendix V.

Water quality throughout the Serpent River Watershed continues to meet and remain well below the assessment criteria established for the protection of aquatic life. Annual average concentrations for all parameters in 2017 were better than assessment criteria at all locations, and cobalt concentrations remained close to or below the detection levels at all stations (Appendix V).

The annual average barium concentration at SR-06 appears elevated (0.606 mg/L) in comparison to other SRWMP stations. However, it is below the assessment criteria of 1.0 mg/L and well below levels considered to be toxic to the aquatic environment (8.0 mg/L; WHO 2001). The elevated level is attributable to the increased barium chloride addition rates that were required to maintain control of radium concentrations upstream at the Stanleigh treatment plant. No acute or chronic toxic effects to aquatic biota were observed upstream at the Stanleigh site's final discharge (CL-06) as a result of increased barium concentrations (RAL, 2018). Response monitoring downstream at the outlet of May Lake (SR-15) continued in 2017 in an effort to determine if there are any impacts to the environment as a result of the elevated barium.

As previously mentioned, according to the most recent Ambient Water Quality Guidelines from the BCMOE, manganese and sulphate are hardness dependent (BCMOE, 2006, 2013). Toxicity studies for both parameters demonstrated amelioration of toxicity with increasing water hardness, and were used to develop new water quality guidelines in the province for these substances. Therefore, based on this information, a specific assessment criterion for sulphate has been established for each station in the SRWMP. For example, sulphate concentrations at SR-08 are higher than other receiving environment stations. However, mean water hardness at this station is 223.9 mg/L thus the resulting criteria for sulphate at this location is 429 mg/L. In 2017, all results at SR-08 fell within BCMOE guidelines for the protection of aquatic life (BCMOE, 2013). A review of the data indicates that sulphate annual concentrations have continued to decrease over the past five years (Figure 3.1.a.). Sulphate assessment criteria for individual stations has been included in Appendix V of this report with the detailed data, as well as in Table B-1, Appendix B, of the *Cycle 4 Study Design for the SRWMP, SAMP, and TOMP* (Minnow, 2016).

Figures 3.1.a to 3.1.c show five-year trends of annual average concentrations for the minerelated parameters sulphate, radium, and uranium at the following key locations:

- SR-01, Quirke Lake Outlet;
- SR-06, McCabe Lake Outlet;
- SR-08, Nordic Lake Outlet:
- DS-18, Halfmoon Lake Outlet.

Based on a review of 5 years of data, annual sulphate concentrations at all key lake outlets are well below the assessment criterion of 128-429 mg/L established for each station. Furthermore, annual concentrations have been gradually decreasing at the McCabe Lake (SR-06) and Nordic Lake (SR-08) stations (Figure 3.1.a).

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With the exception of DS-18, radium concentrations at the remaining three key locations are more than an order of magnitude below the assessment criteria of 1.0 Bq/L, and appear to have reached relatively stable levels based on review of the data (Figure 3.1.b). In 2017, the annual average radium concentration of 0.193 Bq/L at DS-18 is elevated when compared to previous annual concentrations in the last five years (Appendix V). This is likely attributable to a spike observed in August when a heavy rain event may have caused flushing through the historic tailings spill upstream in the Halfmoon wetland. Concentrations are typically significantly lower at this time of year when flow is much lower. However, all concentrations have consistently remained well below 1.0 mg/L (Figure 3.1.b).

Annual uranium concentrations at all four key locations appear to be relatively stable and all values in 2017 remained well below assessment criteria of 0.0150 mg/L (Figure 3.1.c).

Table 3.2 2017 SRWMP Location Annual Average Results Summary

Parameters			рН	SO4 ⁵	Ra(T)	U	Ва	Со	Fe	Mn ⁵	Hardness
				(mg/L)	(Bq/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	(mg/L)	mg/L as CaCO₃
Assessment Criteria ¹	Wetland and lake	benchmarks	6.5	128-429	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland/Stream I	benchmark ²	5.2						1.69		
	Lake benchmark	3							0.49		
MDL ⁴	Lake benefitiark			0.1	0.005	0.0005	0.005	0.0005	0.02	0.005	0.5
Location		# of samples collected									
Reference	Туре										
D-4	Lake	2	6.8	3.5	< 0.007	< 0.0005	0.013	< 0.0005	0.04	0.021	9.6
SR-18	Lake	2	6.8	4.0	< 0.007	< 0.0005	0.043	< 0.0005	0.07	0.025	10.4
SR-19	Lake	4	7.0	3.0	0.008	< 0.0005	0.019	< 0.0005	0.36	0.031	14.4
SR-16	Wetland/Stream	4	5.7	1.1	< 0.007	< 0.0005	0.007	0.0007	0.94	0.038	7.4
SR-17	Wetland/Stream	4	5.8	2.8	0.007	< 0.0005	0.022	0.0008	0.73	0.048	11.8
Near Field											
D-5		4	6.8	11.3	0.040	0.0013	0.045	< 0.0005	0.07	0.026	20.5
D-6		4	6.7	18.8	< 0.007	< 0.0005	0.013	< 0.0005	0.19	0.102	28.7
DS-18		4	6.8	59.8	0.193	0.0008	0.017	< 0.0005	0.60	0.037	83.5
M-01		4	6.8	10.0	0.016	0.0034	0.015	< 0.0005	0.58	0.070	36.3
Q-09		4	6.7	44.8	0.052	0.0015	0.055	< 0.0005	0.17	0.036	55.6
Q-20		1	6.9	19.0	< 0.007	< 0.0005	0.018	< 0.0005	0.04	0.030	37.1
SC-01		1	6.9	16.0	< 0.007	< 0.0005	0.009	< 0.0005	0.07	0.010	26.1
SR-06		4	7.0	33.5	0.089	0.0007	0.606	< 0.0005	0.03	0.011	52.6
Far Field											
SR-15		2	6.9	32.0	0.069	< 0.0005	0.149	< 0.0005	< 0.02	0.005	52.3
SR-01		1	6.9	31.0	0.028	0.0011	0.035	< 0.0005	< 0.02	0.003	38.3
SR-08		4	7.1	150.0	0.026	0.0009	0.017	< 0.0005	0.05	0.036	186.3

Notes:

Bold indicates exceedance of evaluation criteria value

¹ Assessment criteria as per Table 4.5, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland/stream stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

Figure 3.1.a. Annual Average Sulphate Concentrations at SR-01, SR-06, SR-08, and DS-18, 2013-2017

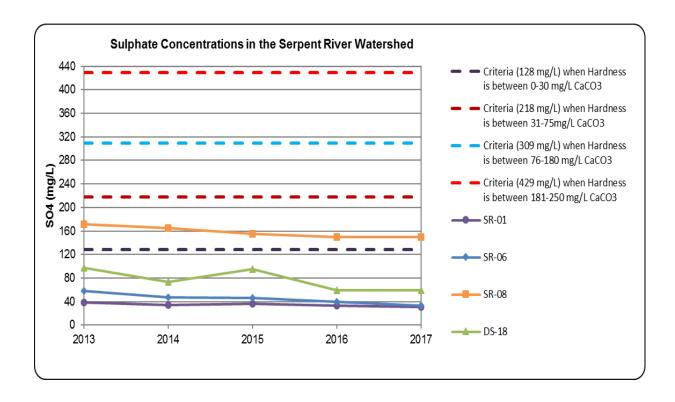


Figure 3.1.b. Annual Average Radium Concentrations at SR-01, SR-06, SR-08, and DS-18, 2013-2017

2017 SRWMP Annual Water Quality Report

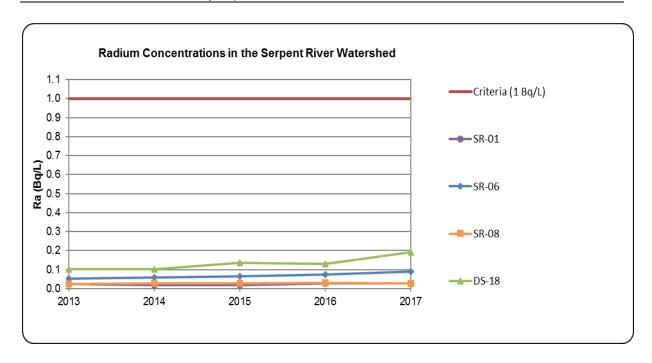
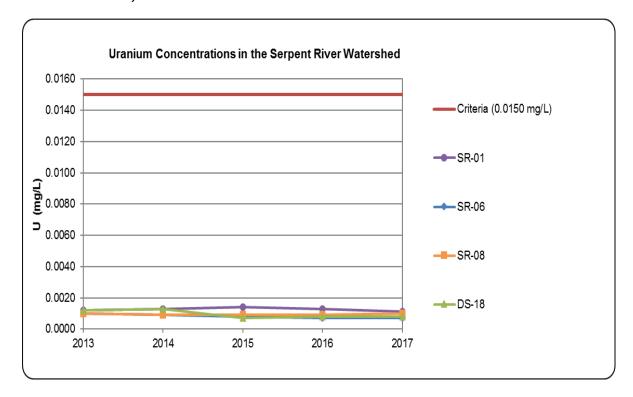


Figure 3.1.c. Annual Average Uranium Concentrations at SR-01, SR-06, SR-08, and DS-18, 2013-2017



4 DISCUSSION

4.1 Response Monitoring

Beginning in 2016, monitoring at the outlet of May Lake (SR-15) was voluntarily re-established in response to elevated barium concentrations upstream at the outlet of McCabe Lake (SR-06); it was previously removed in the SRWMP Cycle 3 Study Design (Minnow 2009). Response monitoring was also initiated in December 2017 due to a non-compliance in the monthly mean radium concentration at the Stanleigh final discharge (CL-06). As a result, supplemental monitoring was implemented at the first receiving station McCabe Lake (SR-06) and near the mixing point with the Serpent River at Pecors Lake (SR-03). This included increased monitoring frequency at SR-06 as well as re-establishing monitoring at SR-03, which had also been removed in the Cycle 3 Study Design (Minnow, 2009). Both locations also included twice-monthly toxicity. Response monitoring at these locations will continue into 2018 until radium control is achieved at the CL-06 final discharge, and concentrations fall below compliance limits. The non-compliance at CL-06 is described in further detail (including all results) in the 2017 Rio Annual OCM Report, while all water quality results from supplemental and regulatory sampling for SR-06 and SR-03 are provided in Appendix V of this report.

4.2 SRWMP Performance Monitoring Program Changes

There were no changes to methodology in 2017.

4.3 Changes to Location Classification and Frequency

Besides the addition of the SRWMP stations SR-15 and SR-03 for the purpose of short-term monitoring, and the increased monitoring frequency at SR-06, there were no other changes to location classification or frequencies in 2017. Following completion of the 2016 Serpent River Watershed Cycle 4 State of the Environment Report, cobalt, iron, and manganese were reintroduced at all SRWMP stations to assist in future interpretations of the loadings of these substances into the watershed (Minnow, 2016). In addition, hardness was added as an ancillary parameter to all SRWMP stations as it assists in interpretation of water quality concentrations for manganese and sulphate, as discussed in the approved *Cycle 4 Study Design for the SRWMP, SAMP and TOMP* (BCMOE, 2006, 2013, and Minnow, 2016). These additions are outlined in Table 2.1.

4.4 Interim Assessment in Support of Representative Public Radiation Dose Estimation

The Canadian Nuclear Safety Commission (CNSC) requested that Rio Algom Limited and Denison Mines Inc. provide annual reporting of the radiation dose to the public associated with the closed uranium mine sites in the Serpent River Watershed. Historically, estimates of the public dose had been based on the use of very conservative values to demonstrate that public dose in the vicinity of Elliot Lake did not exceed the upper dose limit. Measurements of radon and gamma collected during mine operations resulted in dose estimates less than 5% of the annual public dose limit of 1 mSv/a.

However, to determine an updated and more realistic representative annual public dose estimation for a person residing in Elliot Lake, a preliminary design for a monitoring program to support public dose estimation was prepared in early 2016. Details of the design program were provided in the document entitled *Preliminary Design Monitoring Program to Support Public Dose Estimation* (Ecometrix Incorporated (Ecometrix), 2016), which was included as an appendix in last year's *SRWMP Annual Water Quality Report 2016* (RAL, DMI, 2017)

In 2016, components of the design monitoring program were completed. This included quarterly site-specific radiation surveys of public walking trails (for radon and direct Gamma specifically), analysis of radionuclides in drinking water, and a community survey. The community survey was conducted to determine the amount of time a representative person spent hiking on the mining properties as well as information about their consumption of fish on local lakes. Based on the interim public dose calculations using the data collected in 2016, it can be concluded that the public dose to the representative person is approximately 0.012 mSv/a, after correction for background exposure. This interim public dose estimation is intended to provide annual interim dose values until 2019. Details of the interim dose are provided in the document entitled *Interim Public Dose Estimation for the Closed Mines of the Serpent River Watershed* (Ecometrix, 2017), which is included in Appendix VI of this report.

The public dose estimation is expected to be updated as part of the 2019 State of the Environment report when changes to the monitoring data used to calculate the interim dose are projected to include updated sport fish tissue analysis. The annual dose reporting will be based on periodic updates undertaken as part of the five-year SOE Report.

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REFERENCES

- Beak International Incorporated 1999. Serpent River Watershed Monitoring Program Framework Document. February 1999.
- Ecometrix Incorporated, 2016. Preliminary Design for a Monitoring Program to Support Public Dose Estimation, Prepared for Rio Algom and Denison Mines, September, 2016
- Ecometrix Incorporated, 2017. Interim Public Dose Estimation for the Closed Mines of the Serpent River Watershed. February 2018.
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- Minnow Environmental Inc., 2009a. Serpent River Watershed State of the Environment. Prepared for Rio Algom Limited and Denison Mines Inc. January 2009.
- Minnow Environmental Inc., 2009b. Serpent River Watershed Monitoring Program, Cycle 3 Study Design. Prepared for Rio Algom Limited and Denison Mines Inc. May 2009.
- Minnow Environmental Inc., 2009c. Monitoring Framework for Closed Uranium Mines, Near Elliot Lake. Prepared for Rio Algom Limited and Denison Mines Inc. May 2009.
- Minnow Environmental Inc., 2011. Serpent River Watershed State of the Environment Report.

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- Minnow Environmental Inc., 2016b. Serpent River Watershed Cycle 4 State of the Environment Report. Prepared for Rio Algom Limited and Denison Mines Inc. November 2017.
- Rio Algom Limited, 2017. OCM Annual Report. March 2018.
- Rio Algom Limited and Denison Mines Inc., 2017. Serpent River Watershed Monitoring Program 2017 Annual Water Quality Report. March 2018.

APPENDIX I Performance Monitoring Changes 2015 - 2019 Evolution of Programs





March 9, 2016 via e-mail

Karina Lange
Project Officer for Wastes and Decommissioning Division
Canadian Nuclear Safety Commission
280 Slater Street
P.O. Box 1046, Station B
Ottawa, ON, K1P 5S9

Dear Ms. Lange:

Re: Serpent River Watershed Cycle 4 State of the Environment Report

Denison Mines Inc. (DMI) and Rio Algom Limited (RAL) are pleased to submit the Serpent River Cycle 4 State of the Environment (SOE) Report (2010 to 2014). The report presents and integrates the monitoring data obtained through the Elliot Lake closed mines monitoring programs, namely the Serpent River Watershed Monitoring Program (SRWMP), the Source Area Monitoring Program (SAMP) and the TMA Operational Monitoring Program (TOMP). The report covers the period of January 1, 2010 to December 31, 2014 although historical data has been considered for trend analysis.

This report represents the completion of the fourth cycle of the SRWMP. A complete list of all study design and interpretive reports prepared since the start of Cycle 1 is provided in Table 5.1. This table also summarizes the time frame covered for each cycle and the key changes to each of the monitoring programs over time.

We are also distributing this Cycle 4 State of the Environment Report to the members of the Joint Regulatory Review Group (JRG; distribution attached). We look forward to your review of the report and the opportunity to address and any questions or comments you may have.

Yours very truly,

Denison Mines Inc. Rio Algom Limited

lan Ludgate, Manager Debbie Berthelot, Reclamation Manager

cc: Distribution List

Revision 2016.01 Page 1

Table 5.1: Summary of the Elliot Lake monitoring programs; documents produced and changes to the programs during each cycle.

Cycle	Report Title	Year	Period Covered	Description Of Changes To The Monitoring Programs Within Each Cycle					
	Serpent River Watershed Monitoring Program Framework Document.	1999							
	In-Basin Monitoring Program Report	1999	historical monitoring data						
Cycle 1	Serpent River Watershed and In-Basin Monitoring Program – Implementation Document.	1999		SRWMP, IBMP, SAMP and TOMP were developed based on program objectives and existing monitoring data collected over the period of operations and decommissioning.					
	Serpent River Watershed Monitoring Program -1999 Study	2001	1999 - 2000						
	In-Basin Monitoring Program for the Uranium Tailings Areas - 1999 Study.	2001	1999 - 2000						
	Overview of Elliot Lake Monitoring Programs and Source Area Monitoring Program Design.	2002		Changes only SRWMP most associated with optimization after first cycle of program was complete: - monitoring substances reduced to mine indicator parameters (barium, cobalt, DOC, iron, manganese, Ra-226,					
	TMA Operational Monitoring Program Design (TOMP).	2002		selenium, silver, sulphate and uranium),					
010	Cycle 2 Study Design – Serpent River Watershed and In- Basin Monitoring Programs.	2004	2000 2004	- addition of two lake reference stations (Summers and Semiwite lakes) and 3 stream reference areas (SR-16, SR and SR-18); - removal of shallow lakes for sediment and benthic sampling (Westner, Grassy, Halfmoom, Upper Cinder and Hole					
Cycle 2	Serpent River Watershed Monitoring Program: Cycle 2 Interpreative Report	2005	2000 -2004	lakes); - removal of some stream sediment and benthic stations (D-15, SC-03 and SR-07);					
	Serpent River In-Basin Monitoring Program: Cycle 2 Interpretive Report - 2004 Study.	2005		 removal of Depot Lake and Serpent Harbour; addition of May Lake; the transfer of some SRWMP stations to SAMP or TOMP (N-12, ECA-131, P-11, MPE and Q-23); fish health assessment eliminated based on performance, fish community assessment added for McCabe Lake and 					
	Serpent River Watershed State of the Environment	2009		fish tissue monitoring reduced in scope based on performance.					
	Monitoring Framework For Closed Uranium Mines Near Elliot Lake	2009		IBMP eliminated based on objectives of program being achieved. SAMP and TOMP:					
	In Basin Monitoring Program, Cycle 3 Study Design	2009		- removal of silver, selenium based on performance and removal of conductivity based on redundancy with sulphate - DOC, hardness and flow added at selected stations.					
Cycle 3	Serpent River Watershed Monitoring Program: Cycle 3 Study Design	2009	2005- 2009	SRWMP: - removal of selenium and sliver based on performance,					
Cycle 3	Source Area Monitoring Program Revised Study Design.	2009	2003- 2009	- removal of station SR-12, ELO, SR-09, SR-15, SR-02, SR-03, SR-11, P-01, QL-01 and SR-16 and SR-17 based on performance;					
	Tailing Management Area Monitoring Program (TOMP) Revised Study Design	2009		- monthly monitoring frequency reduced to quarterly; - sediment and benthic monitoring removed from Whiskey, Evans and Cinder Lakes based on redundancy,					
	Serpent River Watershed State of the Environment Report.	2011		 depositional streams (Q-20, D-6, SR-06, M-01 and SR-08) based on very high natural variability masking results; fishing in McCabe Lake and fish tissue monitoring eliminated based on performance. 					
Cuala 4	Cycle 4 Study Design For the SRWMP, SAMP and TOMP.	2014 ^a	2010 2014	Minor changes to SAMP and TOMP. SRWMP: - elimination of reference stations SR-05, P-222 and SR-14; removal of cobalt as substance for manifering, addition of DOC:					
Cycle 4	Serpent River Watershed Cycle 4 State of the Environment	2016	- 2010 - 2014	 removal of cobalt as substance for monitoring, addition of DOC; far-field lakes removed from the program (Hough, Pecors and McCarthy); removal of Rochester Lake as a sediment and benthic reference area; reduction in benthic and sediment sampling to 1/10 years based on measured deposition rates. 					

^a Study Design was submitted to CNSC and JRG in 2014 but reissued with agency comments in 2016.

APPENDIX II Flagged Data Results



Annual SRWMP Data Flags 2017 Revision 2015-01



Issued on: June 30, 2015

Expires: June 30, 2020

Report Form: RC8.7.3.01

Location	Analyte	Date	Low	Hi	Result	Comment
DS-18	Mn	2017-02-23	0	0.062	0.098 mg/L	Result is a 14-year high confirmed by repeat analysis but still consistent with historic values, will continue to monitor at the current quarterly frequency.
M-01	U	2017-02-23	0.0011	0.0046	0.0053 mg/L	Result is slightly above the high flag limit but still consistent with previous values at this location.
SR-08	Mn	2017-02-23	0.019	0.051	0.053 mg/L	Result is slightly above the high flag limit but still consistent with previous values at this location.
SR-15	SO4	2017-05-26	34.4	38.6	31.0 mg/L	Result is slightly below the low flag limit but consistent with decreasing trends observed upstream at SR-06.
SR-17	pHF	2017-05-24	5.1	6.1	6.2	Result is a historic high confirmed by a repeat measurement but is only slightly above the high flag limit, will continue to monitor at the current quarterly frequency.
D-4	Ва	2017-11-21	0.013	0.013	0.012 mg/L	Result is slightly above the high flag limit but consistent with previous values at this location in the last five years.



Annual SRWMP Data Flags 2017 Revision 2015-01



Issued on: June 30, 2015

Expires: June 30, 2020

Report Form: RC8.7.3.01

Location	Analyte	Date	Low	Hi	Result	Comment
Q-20	SO4	2017-11-21	19.9	23.6	19.0 mg/L	Result is a 10-year low but consistent with a gradually decreasing trend and only slightly below the low flag limit. Will continue to monitor at the current annual frequency.

APPENDIX III Laboratory QA/QC Results



DEN-ANN17 REPORT CODE:

Annual Quality Assessment Report REPORT TITLE:

1.0 REVISION:

ISSUED BY:

Quality Coordinator

AUTHORIZED BY: ROLLINGAI Manager

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BACKGROUND

SGS Environmental Services entered into an agreement with Denison Environmental Services for the analytical lab to provide analysis according to RFT #05-016. Below is a summary of the laboratory quality management system, key actions taken by the laboratory for samples analyzed during 2017, as well as a summary of the significant findings and the corrective actions implemented.

2. QUALITY MANAGEMENT SYSTEM

SGS Environmental Services is accredited to ISO/IEC 17025 by the Canadian Association for Laboratory Accreditation (CALA), for specific tests listed in the scope of accreditation. ISO/IEC 17025 addresses both quality management and the technical aspects of operating a testing laboratory. he quality management system at SGS Environmental Services consists of a documented quality system, which is manual, methods, written instructions, standard operating procedures, and data approval criteria) is in place and includes directed by the Quality Coordinator who is independent of the production area. All appropriate documentation (quality both general and method specific quality control requirements.

preparation control samples, certified reference material analysis, and instrument control samples, as appropriate for the individual methods. Matrix matching of reference materials to samples is always attempted. Frequency of insertion of control samples is method specific and follows legislated guidelines. A summary of the quality control recoveries is The quality control procedures include duplicate samples, spiked blanks, spiked replicates, reagent/instrument blanks, presented in the tables following.

3. QUALITY CONTROL PARAMETERS

replicates, reagent/instrument blanks, preparation control samples, certified reference material analysis, and instrument control samples, as appropriate for the individual methods. Matrix matching of reference materials to samples is always The analysis of quality control samples is method specific and includes duplicate samples, spiked blanks, spiked attempted. Frequency of insertion of control samples is method specific and follows legislated guidelines and/or customer All GC analyses for Denison Environmental Services are tracked in unique files, specific to Denison Environmental Services. The samples are processed as part of our "worksheet" batch system and a compilation of all Denison Environmental Services QC data for the parameters tested during 2017 has been compiled below. requirements.



MAJOR ACHIEVEMENTS IN QUALITY CONTROL

4

- SGS Environmental Services performed 7769 analyses with 6685 QC checks, which represents 86% QC for sample analysis. This level of QC is significantly higher than the lab standard, which is generally 20%.
- All blank data results were within the data quality objectives. Corrective Action: N/A
- All CRM data results were within the data quality objectives. Corrective Action: N/A
- No duplicate value exceeded the data quality objectives. Corrective Action: N/A
- No spike blanks exceeded the data quality objectives. Corrective Action: N/A
- Barium (Ba) reporting limit as per the Denison reports is set at 0.005 mg/L. The detection limit normally reported for Ba by SGS Environmental Services - Lakefield is 0.0001 mg/L. The concentration of the Ba spike is 0.005 mg/L. The Ba spikes all passed under lab limits; however, with the detection limit set at 0.005 mg/L for the Denison reports a number of the spikes appear to fail when taking into account the detection level. **Corrective**
- fron (Fe) reporting limit as per Denison reports is set at 0.02 mg/L. The detection limit normally reported for Fe by SGS Environmental Services Lakefield is 0.007 mg/L. The concentration of the Fe spike is 0.01 mg/L. The Fe spikes all passed under lab limits; however, with the detection limit set at 0.02 mg/L for the Denison reports a number of the spikes appear to fail when taking into account the detection level. Corrective Action: N/A
- No spike duplicates fell outside of the data quality objectives. Corrective Action: N/A

GC DATA SUMMARY

5.1. Blank Data

Parameter	Unit	Required	Mean Blank Reguli	Number of Blanks	Number Greater Ihan Umit	Number Outside +/- Limit
Acidity	mg/L as	-	5,1	45	97	<u>£</u>
Alkalinity	mg/L as cacos	2	0.88	0		
Barium	mg/L	0.005	0.023	192		



Cobalt	mg/L	0.0005	0.0002	156		-
Dissolved Organic Carbon	mg/L	6.5	0.25	Ø		
lron	mg/L	0.02	0.009	223	7	
Hardness	mg/L as CaCO3	0.5	0.22	62		
Manganese	mg/L	0.002	0.0009	166		
Sulphate	Пд/Г	0.1	0.053	187		i i
Total Dissolved Solids	mg/L	- P	15	11	11	
Total Suspended Solids	mg/L	-	0.48	364		
Uranium	mg/L	0.0005	0.0002	153	0	0
Zinc	ma/L	0.001	0,001			

5.2. CRM Data

Parameter	Unit	Certified	Lower Limit (at 10% Rel. Error)	Upper Limit (at 10% Ref Error	Number of CRM's	Mean Value	% RSD (Precision)	Number OC CRM outside of DGO
Acidity	mg/L as	90	45	53	26	49.9	2.85	
Alkalinity	mg/L as CaCO3	47.2	42.5	51.9	o	48.0	2.09	
Barium	mg/L	0.1	0.09	0,11	185	0.10	3,05	
Cobalt	mg/L	0.1	60:0	0.11	<u>8</u>	0,10	2.49	
Iron	mg/L	500	450	550	210	0.102	7.41	14
Hardness	mg/L as CaCO3	0.1	0	0	47	78.0	10. 12.	
Manganese	mg/L	0.1	60.0	0.71	2	0.102	2.53	
Sulphate	mg/L	ĸ	4.5	5.5	179	4.82	1.96	
Total Suspended Solids	mg/L	901	06	110	348	97.8	2.17	
Uranium	ma/L	0.1	0.09	0.11	156	0.100	3.44	



5.3. Duplicate Data

Parameter	Unit	Expected Recovery (Ref. %)	Lower Umit (Rel. %)	Upper Limit (Rel. %)	Number of Duplicates	Meen Recovery (%)
Acidity	mg/L as CaCO3	100	8	110	95	88
Silver	mg/L	100	8	110	Constant Constant	100.0
Alkalinity	mg/L as CaCO3	100	8	110	60	99.1
Arsenic	mg/L	9 <u></u>	8	110	-	0.66
Barlum	mg/L	100	06	110	189	97.0
Cobalt	mg/L	901	06	110	152	0.96
Copper	mg/L	\$	90	110	-	0.76
200	тgЛ	100	8	110	-	98.9
Iron	mg/L	100	06	110	201	97.2
Hardness	mg/L as CaCO3	100	90	110	18	88.4
Manganese	mg/L	100	06	110	161	97.1
Nickel	Πg/L	100	06	110	-	95.0
Lead	mg/L	100	90	110	1	94.0
Selenium	mg/L	100	06	110	1	97.0
Sulphate	mg/L	100	8	110	138	97.2
Total Dissolved Solids	mo/l	100	06	110	=	97.8
Total Suspended Solids	molt	100	8	110	352	97.4
Uranium	ma/l.	100	06	110	154	98.1



5.4. Spike Blank Data

Number Spike Blank within \$95 OC protocols and below Denison MDL	0	0	0	0	٥	0	0	0	0	o	٥	0	0	0	0	0
Number Spike Blank outside of DGO	0	0	o	0	173	0	0	25	0	-	5	0	0	0	2	0
Precision (%RSD)	72		5.4		18.0	1.4		39.9	2.2	23.3	14.1			1,6	7,8	5.7
Mean % Recovery	112.6	75.0	92.6	103.1	53.7	0.66	0.66	117.8	78.2	107.0	111.5	104.0	106.3	121.9	99.2	39.0
Number of Spike Blank's	100		6	-	193	158	2	179	7	34	153	-	-	187	114	47
Upper limit (at 30% Rel. Error)	13	0.00026	12.22	0.00832	0.00832	0.0026	5	0.013	0.065	o	0.004	0.0065	0.00104	5.2	0.00104	0.0078
Lower Limit (at 30% Rei. Error)	7	0.00014	6.58	0.00448	0.0035	0.0014	7	0.007	0.035	0	0.0021	0.0035	0.00056	2.8	0.00056	0.0042
Certified Value	10	0.0002	4.6	0.0064	0.005	0.002	10	0.01	0.05	0	0.003	0,005	0.0008	4	0.0008	0.006
Vait	mg/L as CaCO3	mg/L	mg/L as caco3	mg/L	mgvL					- 3	mg/L	mg/L	ШGL	mg/L	mgvL	marl
Perameter	Acidity	Silver	Alkalinity	Arsenic	Barium	Cobalt	200	Iron ICP-MS	ron ICP. OES	Hardness	Manganese	Nickel	Selenium	Sulphate	Uranium	Zinc



5.5. Spike Duplicate Data

Parameter	Unit	Certified Value	Lower Limit (at 30% Ret. Error)	Upper (imit (at 30% Rel. Error)	Number of Spike Rep's	Mean % Recovery	Precision (%RSD)	Number Spike Rep autside of DGO	Number Spike Dup. outside of SGS QC protocols and below Denlson MDL
Arsenic	mg/L	0.64	0.448	0.832	20	110.7	9.20	0	0
Ватіит	mg/L	4	2.8	5.2	153	39.2	0.58	တ	0
Cobalt	шg/L	2	1,4	2.6	52	93.5	0.31	9	0
DOC-Low	mg/L	100	02	130	30	73.7	17.6	0	0
DOC-SK	mg/L	100	02	130	24	0.66	3.93	0	0
Iron ICP-MS		0.05	0.035	0.065	9	93.0	0.01	ev	0
Iron ICP. OES	- 1	0.05	0.035	0.065	_	104.0		o	0
Hardness		4.55	3.18	5.92	51	48.4	9.41	o	0
Manganese	mg/L	3.2	2.24	4.15	22	102.6	0.52	en	0
Lead	mg/L	3.2	2.24	4,16	1	106.5		0	0
Selenium	mg/L	9.0	0,56	1.04	20	106.7	8.29	0	0
Sulphate	mg/L	100	70	130	8	97.3	5,73	0	0
Uranium	mg/L	0.8	0.56	1.04	128	97.5	60.0	0	0



5.6. QC Frequency

Total Analysis:	7769
Sum of QC Insertion:	6685
Total Number of Spike Duplicates:	834
Total Number of Spike Blanks:	1187
Total Number of Duplicates:	1486
Total Number of CRM:	1548
Total Number of Blanks:	1630

6. CONCLUSION & SIGNIFICANT FINDINGS

SGS Environmental Services analyzed QC samples for this project beyond the lab standard of 20% QC insertion. Where the data quality objectives for the lab were exceeded, the additional QC samples analyzed within the run supported the data values and data was released on this basis.

SGS Environmental Services remains committed to delivering data that meets and/or exceeds the data quality objectives for Denison Environmental Services and staff will continue to work closely with Denison Environmental Services staff to ensure all objectives are achieved in 2018.

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Environmental Monitoring and Reclamation Research

An ISO:17025 Accredited Laboratory

²²⁶Ra DATA QUALITY REPORT

2017 Annual

Prepared by:

Troy Maki - Analytical Chemist

Troy Maki

Reviewed by:

Dr. Graeme Spiers - Director

Elliot Lake Research Field Station of Laurentian University Date: January 19, 2018

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Environmental Monitoring and Reclamation Research

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1 Background

Elliot Lake Research Field Station (ELRFS) entered into an agreement with Denison Environmental Services (DES) for the analytical laboratory to provide ²²⁶Ra analysis according to the ELFRS Offer of Services document submitted to DES on December 3, 2010. Please find below the summaries of the 2017 annual Quality Control (QC) results for blanks, duplicates, certified reference material (CRM), and spiked sample analysis.

The Analytical Services Laboratory of the Elliot Lake Research Field Station (ELRFS) was established in 1992. The initial work of the laboratory was to support research into the effects of low-level radioactivity on the environment resulting from regional uranium mining activities.

From this base, the laboratory has provided analytical services in support of local decommissioning and environmental monitoring programs, and in support of academic research. While the laboratory specializes in radionuclide analysis, it also provides a wide range of inorganic services for environmental samples, including solid wastes, effluents, receiving waters, ground waters, soils, sediments, geological materials, plant tissues and animal and fish tissues. The ELRFS analytical team will also complete specialty analyses outside of the scope of accreditation, following good laboratory practice procedures, using similar QA/QC protocols.

2 Quality Management System

ELRFS is ISO/IEC 17025:2005 accredited by the Canadian Association of Laboratory Accreditation (CALA) for specific environmental tests listed in the Scope of Accreditation. Accreditation is the formal recognition of the competence of a laboratory to achieve and demonstrate the highest levels of scientific and management excellence through the combined principles of Competence, Consistency, Credibility and Communication.

The quality management system at ELRFS consists of a documented quality system stating the quality policy, quality system and quality practices designed to demonstrate quality control operations are being carried out, to ensure accountability of data, to assure traceability of reported data, and to show that reasonable precautions are being taken against the possibility of falsification of data. Within this manual, Quality Assurance Procedures and Standard Operation Procedures define the laboratory operational duties that guide the analytical QC data. This includes a minimum target of 20% of the samples analysed being distributed as blanks, duplicate analysis, CRMs, and spiked samples. The sample and QC results are logged into excel spread sheets and Envista data management systems with monthly and annual QC reports generated.



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3 Quality Control Parameters

All QC parameters are taken directly from the Excel spread sheets and Envista. DES samples are processed as part of the worksheet batch system. A compilation of all QC data appropriate to the parameters tested has been compiled below.

The QC summary reports are presented as control charts with the mean +/- 1 standard deviation illustrated as the SD Level, the mean +/- 2 standard deviations illustrated as the Warning Level and the mean +/- 3 standard deviations illustrated as the Control Level.

Control Level - If the Control Level is exceeded, the analysis of standards and samples must be repeated and if the repeat analysis exceeds the Control Level again, corrective action is required.

Warning Level – If 2 or more consecutive points exceed the Warning Level, another standard must be analyzed and if this analysis exceeds the Warning Level again, corrective action is required.

SD Level – If 4 consecutive results exceed the SD Level, analyse the next sample and if the SD Level is exceeded again, corrective action is required.

4 Notable Occurrences /Actions

Through the year of 2017, ELRFS analyzed 101 batches totaling 1197 samples for ²²⁶Ra. Each batch incorporated blank, CRM, duplicate, and spiked samples providing greater than 25% quality control samples. All quality control samples are within control limits (mean +/- 3SD).

Twenty-four quality control samples exceeded the warning (mean +/- 2SD) levels. This included five QC Blank (Figure 1a) samples, 4 QC Duplicate (Figure 2) samples, four QC CRM (Figure 3a) samples and 11 QC Spike (Figure 4a) samples. All samples exceeding the warning level were not consecutive, with the next consecutive QC sample falling within the warning level (mean +/- 2SD) limit, thus no corrective actions were required. No QC samples exceeded objectives.

5 QC Data Summary

Table 1. Summary of QC results for 2017.

Quality Element	Unit	Objective	Total Number of Samples QC Samples	Expected Value	Mean	Number Outside Warning Limit	Number Outside Control Limit	Number Exceeding Objective
Blank	Bq/L	0.01	101	0	0.00062	5	0	0
Duplicate % error	%	20	101	0	6.31	4	0	0
CRM	Bq/L	20	101	0.044	0.044	4	0	0
Spike	Bq/L	20	101	0.249	0.251	11	0	0



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1 Blanks

The blank sample is composed of ultra pure water and is treated in an identical manner, including all of the added reagents, as normal samples. The criterion of the blank sample is 0.01 Bq/L which is equal to 6 counts per 100 min (0.06 cpm). The 2017 mean blank value is 1.01 counts per 100min (0.00062 Bq/L). ELRFS uses counts to monitor the blank quality control data.

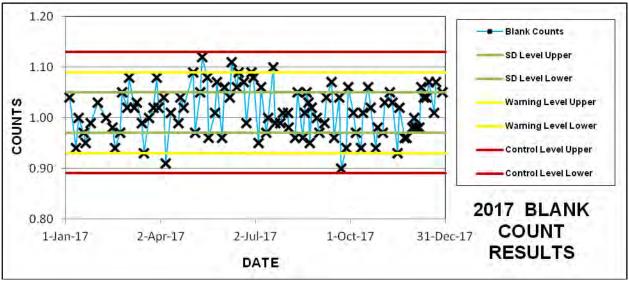


Figure 1a: Blank quality control results for the 2017 year.

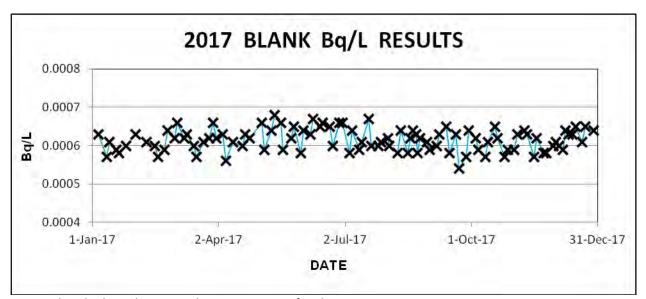


Figure 1b: Blank quality control concentrations for the 2017 year. Note maximum concentrations are 10 times lower than the 0.01 Bq/L criteria.



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2 Duplicates

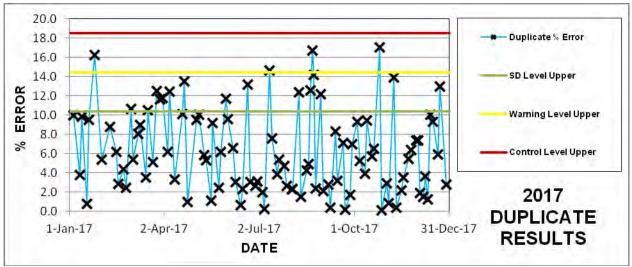


Figure 2: Duplicate quality control results for the 2017 year.

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3 CRM

The CRM material used is ERA # RAD-A and contains 0.044 Bq/L.

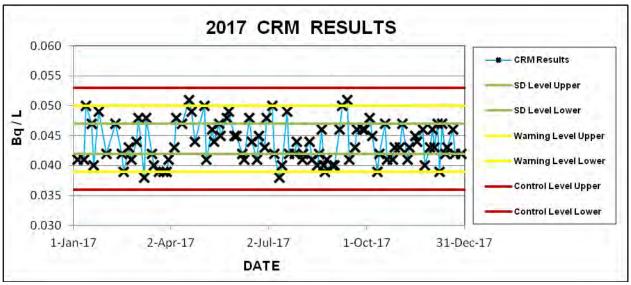


Figure 3a: CRM quality control results for the 2017 year.

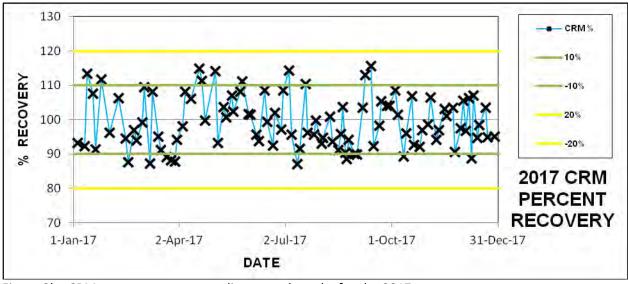


Figure 3b: CRM percent recovery quality control results for the 2017 year.

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4 Spikes

The spike recovery concentration is 0.249 Bq/L.

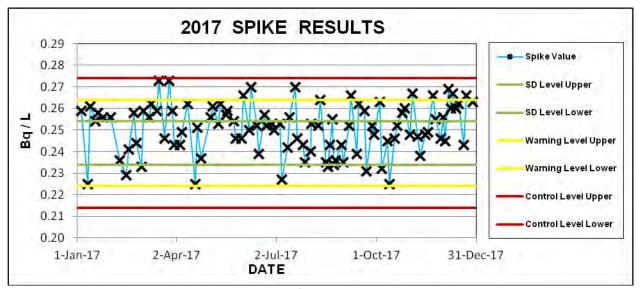


Figure 4a: Spike recovery quality control results for the 2017 year

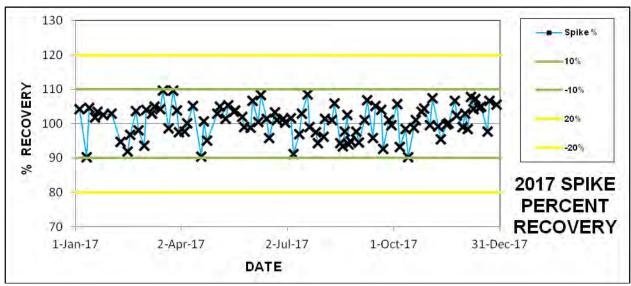


Figure 4b: Percent spike recovery quality control results for the 2017 year.



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QC Frequency

Through the 2017 year, ELRFS analyzed 101 batches totaling 1197 samples for ²²⁶Ra. Each batch incorporated blank, CRM, duplicate, and spiked samples providing greater than 25% quality control samples.

APPENDIX IV Field QA/QC Results



SRWMP DATA QUALITY REPORTING Field Blank Revision: 2016-01



Page 1 of 1

Report Form: RC8.5.4.01b

Date		рН	Sulphate	Radium (total)	Uranium	Barium	Cobalt	Iron	Manganese
			mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	mg/L
SRWI	MP ¹	1	0.2	0.01	0.001	0.01	0.001	0.04	0.004
2017.05	FBD2	5.2	< 0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002
2017.05	FBR5	5.5	< 0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002
2017.11	FBD2	5.2	< 0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002
2017.11	FBR5	5.2	< 0.1	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002
Count		4	4	4	4	4	4	4	4
# Exceedance	es	4	0	0	0	0	0	0	0
Average		5.3	< 0.10	< 0.007	< 0.0005	< 0.0050	< 0.0005	< 0.02	< 0.0020
Max		5.5	< 0.10	< 0.007	< 0.0005	< 0.0050	< 0.0005	< 0.02	< 0.0020
Min		5.2	< 0.10	< 0.007	< 0.0005	< 0.0050	< 0.0005	< 0.02	< 0.0020

¹ SRWMP Field Blank criteria as per Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016) Bold indicates an exceedance in the Field Blank criteria

Valid from: April 29, 2016 Valid to: April 29, 2020



Registry: RC8.5.4.01a

SRWMP DATA QUALITY REPORTING Field Precision Revision: 2016-01



Page 1 of 1

Month	Sample	рН	Sulphate	Radium (T)	Uranium	Barium	Cobalt	Iron	Manganese
			mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	mg/L
2017-05	BSD2	6.6	22.0	< 0.007	< 0.0005	0.013	< 0.0005	0.18	0.117
	D-6	6.6	21.0	< 0.007	< 0.0005	0.013	< 0.0005	0.17	0.111
	variance	0.0%	4.7%	0.0%	0.0%	0.0%	0.0%	5.7%	5.3%
2017-05	BSR5	6.9	9.1	0.018	0.0019	0.016	< 0.0005	0.39	0.095
	M-01	6.9	9.8	0.016	0.0020	0.016	< 0.0005	0.40	0.094
	variance	0.0%	7.4%	11.8%	5.1%	0.0%	0.0%	2.5%	1.1%
2017-11	BSR5	6.8	9.2	0.010	0.0043	0.013	< 0.0005	0.25	0.033
	M-01	6.8	9.3	0.010	0.0041	0.012	< 0.0005	0.23	0.029
	variance	0.0%	1.1%	0.0%	4.8%	8.0%	0.0%	8.3%	12.9%
2017-11	BSD2	6.7	14.0	< 0.007	< 0.0005	0.012	< 0.0005	0.16	0.080
	D-6	6.7	14.0	< 0.007	< 0.0005	0.012	< 0.0005	0.15	0.079
	variance	0.0%	0.0%	0.0%	0.0%	0.0%	0.0%	6.5%	1.3%
Count		4	4	4	4	4	4	4	4
Average		0.0%	3.3%	2.9%	2.5%	2.0%	0.0%	5.8%	5.1%
Max		0.0%	7.4%	11.8%	5.1%	8.0%	0.0%	8.3%	12.9%
Min		0.0%	0.0%	0.0%	0.0%	0.0%	0.0%	2.5%	1.1%
SRWMP ¹ Ta	arget	20%	20%	20%	20%	20%	20%	20%	20%
# Exceedan	ces	0	0	0	0	0	0	0	0

Field Precision criteria as per Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016) Bold indicates an exceedance in the field precision criteria

APPENDIX V Location Results

BSD2

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L
2017-05	32.0	6.6	22.0	< 0.007	0.013	< 0.0005	0.18	0.117
2017-11	22.4	6.7	14.0	<0.007	0.012	<0.0005	0.16	0.080
Count	2	2	2	2	2	2	2	2
High	32.0	6.7	22.0	< 0.007	0.013	< 0.0005	0.18	0.117
Low	22.4	6.6	14.0	< 0.007	0.012	< 0.0005	0.16	0.080
Mean	27.2	6.7	18.0	<0.007	0.013	<0.0005	0.17	0.099
High Limit		8.5	309	1.000	1.000	0.0025	0.49	0.800
Low Limit	0	6.5	0	0	0	0	0	0
Lim Ex	0	0	0	0	0	0	0	0
Frequency	0%	0%	0%	0%	0%	0%	0%	0%
10x Lim Ex	0	0	0	0	0	0	0	0
Frequency	0%	0%	0%	0%	0%	0%	0%	0%

Month	U
	mg/L
2017-05	<0.0005
2017-11	< 0.0005
Count	2
High	< 0.0005
Low	<0.0005
Mean	<0.0005
Mean	<0.0003
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

BSR5

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	35.9	6.9	9.1	0.018	0.016	<0.0005	0.39	0.095	
2017-11	33.0	6.8	9.2	0.010	0.013	<0.0005	0.25	0.033	
Count	2	2	2	2	2	2	2	2	
High	35.9	6.9	9.2	0.018	0.016	< 0.0005	0.39	0.095	
Low	33.0	6.8	9.1	0.010	0.013	< 0.0005	0.25	0.033	
Mean	34.5	6.8	9.1	0.014	0.014	<0.0005	0.32	0.064	
High Limit		8.5	218.0	1.000	1.000	0.0025	0.49	0.800	
Low Limit		5.2							
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	0.0019
2017-11	0.0043
Count	2
High	0.0043
Low	0.0019
Mean	0.0031
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

D-4 Dunlop Lake Outlet

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	10.1	6.7	3.6	< 0.007	0.013	<0.0005	0.04	0.013	
2017-11	9.2	6.8	3.5	<0.007	0.012	<0.0005	0.05	0.028	
Count	2	2	2	2	2	2	2	2	
High	10.1	6.8	3.6	< 0.007	0.013	< 0.0005	0.05	0.028	
Low	9.2	6.7	3.5	< 0.007	0.012	< 0.0005	0.04	0.013	
Mean	9.6	6.8	3.5	< 0.007	0.013	<0.0005	0.04	0.021	
High Limit Low Limit		8.5 6.5	128.0	1.000	1.000	0.0025	0.49	0.800	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	<0.0005
2017-11	<0.0005
0	•
Count	2
High	<0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

D-5 Serpent R. between Denison and Quirke TMAs

Month	FLOW	hard	pHF	SO4	Ra	Ва	Co	Fe	
	L/s	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	
2017-02	1520.00	21.6	6.9	11.0	0.019	0.025	<0.0005	0.05	
2017-05	1928.00	20.5	6.8	11.0	0.036	0.042	< 0.0005	0.05	
2017-08	1379.00	22.7	6.6	14.0	0.094	0.086	< 0.0005	0.09	
2017-11	4843.00	17.2	6.8	9.0	0.010	0.026	< 0.0005	0.06	
Count	4	4	4	4	4	4	4	4	
High	4843.00	22.7	6.9	14.0	0.094	0.086	< 0.0005	0.09	
Low	1379.00	17.2	6.6	9.0	0.010	0.025	< 0.0005	0.05	
Mean	2417.50	20.5	6.8	11.3	0.040	0.045	< 0.0005	0.07	
High Limit Low Limit			8.5 6.5	128.0	1.000	1.000	0.0025	0.49	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Mn	U
	mg/L	mg/L
2017-02	0.024	0.0011
2017-05	0.020	0.0013
2017-08	0.031	0.0017
2017-11	0.029	0.0010
0	4	4
Count	4	4
High	0.031	0.0017
Low	0.020	0.0010
Mean	0.026	0.0013
High Limit	0.800	0.0150
Lim Ex	0	0
Frequency	0%	0%
10x Lim Ex	0	0
Frequency	0%	0%

D-6 Cinder Lake Outlet

Month	FLOW	hard	pHF	SO4	Ra	Ва	Co	Fe	
	L/s	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	
2017-02	159.00	24.0	7.0	14.0	< 0.007	0.015	<0.0005	0.16	
2017-05	89.70	33.2	6.6	21.0	< 0.007	0.013	< 0.0005	0.17	
2017-08	27.80	35.2	6.6	26.0	< 0.007	0.011	< 0.0005	0.30	
2017-11	331.00	22.5	6.7	14.0	<0.007	0.012	< 0.0005	0.15	
Count	4	4	4	4	4	4	4	4	
High	331.00	35.2	7.0	26.0	< 0.007	0.015	< 0.0005	0.30	
Low	27.80	22.5	6.6	14.0	< 0.007	0.011	< 0.0005	0.15	
Mean	151.88	28.7	6.7	18.8	<0.007	0.013	< 0.0005	0.19	
High Limit			8.5	309	1.000	1.000	0.0025	0.49	
Low Limit Lim Ex	0	0	6.5 0	0	0	0	0	0	
		_	_		0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Mn	U
	mg/L	mg/L
2017-02	0.086	<0.0005
2017-05	0.111	< 0.0005
2017-08	0.131	< 0.0005
2017-11	0.079	< 0.0005
Count	4	4
High	0.131	< 0.0005
Low	0.079	< 0.0005
Mean	0.102	<0.0005
High Limit	0.800	0.0150
Lim Ex	0	0
Frequency	0%	0%
10x Lim Éx	0	0
Frequency	0%	0%

DS-18 Halfmoon Lake Outlet

Month	FLOW	hard	pHF	SO4	Ra	Ва	Co	Fe	
	L/s	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	
2017-02		101.0	6.7	73.0	0.203	0.019	0.0005	1.44	
2017-05	232.00	82.3	7.0	59.0	0.188	0.018	< 0.0005	0.31	
2017-08	361.00	61.8	6.8	38.0	0.193	0.015	< 0.0005	0.41	
2017-10	423.00	88.8	6.8	69.0	0.186	0.017	< 0.0005	0.24	
Count	4	4	4	4	4	4	4	4	
High	423.00	101.0	7.0	73.0	0.203	0.019	0.0005	1.44	
Low	232.00	61.8	6.7	38.0	0.186	0.015	< 0.0005	0.24	
Mean	338.67	83.5	6.8	59.8	0.193	0.017	0.0005	0.60	
High Limit			8.5	309	1.000	1.000	0.0025	1.69	
Low Limit			5.2						
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Mn	U
	mg/L	mg/L
2017-02	0.098	0.0013
2017-05	0.014	< 0.0005
2017-08	0.024	0.0008
2017-10	0.013	0.0008
Count	4	4
High	0.098	0.0013
Low	0.013	< 0.0005
Mean	0.037	0.0008
		0.04=-
High Limit	0.800	0.0150
Lim Ex	0	0
Frequency	0%	0%
10x Lim Ex	0	0
Frequency	0%	0%

FBD2

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	<0.5	5.2	<0.1	< 0.007	< 0.005	< 0.0005	< 0.02	< 0.002	
2017-11	<0.5	5.2	<0.1	<0.007	<0.005	<0.0005	< 0.02	<0.002	
Count	2	2	2	2	2	2	2	2	
High	< 0.5	5.2	<0.1	< 0.007	< 0.005	< 0.0005	< 0.02	< 0.002	
Low	< 0.5	5.2	<0.1	< 0.007	< 0.005	< 0.0005	< 0.02	< 0.002	
Mean	<0.5	5.2	<0.1	<0.007	<0.005	<0.0005	< 0.02	<0.002	
High Limit Low Limit		8.5 6.5	309	1.000	1.000	0.0025	0.49	0.800	
Lim Ex	0	2	0	0	0	0	0	0	
Frequency	0%	100%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	<0.0005
2017-11	<0.0005
0	•
Count	2
High	<0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

FBR5

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	<0.5	5.5	<0.1	< 0.007	<0.005	<0.0005	<0.02	< 0.002	
2017-11	<0.5	5.2	<0.1	<0.007	<0.005	<0.0005	< 0.02	<0.002	
Count	2	2	2	2	2	2	2	2	
High	<0.5	5.5	<0.1	< 0.007	< 0.005	< 0.0005	< 0.02	< 0.002	
Low	< 0.5	5.2	<0.1	< 0.007	< 0.005	< 0.0005	< 0.02	< 0.002	
Mean	<0.5	5.3	<0.1	<0.007	<0.005	<0.0005	< 0.02	<0.002	
High Limit Low Limit		8.5 5.2	218	1.000	1.000	0.0025	1.69	0.800	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	<0.0005
2017-11	<0.0005
0	•
Count	2
High	<0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

M-01 Sherriff Creek @ Hwy 108

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-02	34.7	6.6	13.0	0.012	0.015	<0.0005	0.63	0.076	
2017-05	36.6	6.9	9.8	0.016	0.016	< 0.0005	0.40	0.094	
2017-08	43.1	6.9	8.0	0.026	0.018	< 0.0005	1.04	0.081	
2017-11	30.6	6.8	9.3	0.010	0.012	<0.0005	0.23	0.029	
Count	4	4	4	4	4	4	4	4	
High	43.1	6.9	13.0	0.026	0.018	< 0.0005	1.04	0.094	
Low	30.6	6.6	8.0	0.010	0.012	< 0.0005	0.23	0.029	
Mean	36.3	6.8	10.0	0.016	0.015	<0.0005	0.58	0.070	
High Limit Low Limit		8.5 5.2	218	1.000	1.000	0.0025	1.69	0.800	
Lim Ex	0	0	0	0	0	0	2	0	
Frequency	0%	0%	0%	0%	0%	0%	50%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-02	0.0053
2017-05	0.0020
2017-08	0.0021
2017-11	0.0041
Count	4
	0.0053
High	
Low	0.0020
Mean	0.0034
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

Q-09 Serpent R. below Quirke TMA Effluent

Month	FLOW	hard	pHF	SO4	Ra	Ва	Co	Fe	
	L/s	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	
2017-02	1630.00	74.3	6.8	61.0	0.022	0.028	< 0.0005	0.19	
2017-05	2028.00	46.3	6.7	35.0	0.059	0.056	< 0.0005	0.15	
2017-08	1489.00	66.7	6.6	57.0	0.100	0.108	< 0.0005	0.18	
2017-11	4978.00	35.1	6.8	26.0	0.027	0.028	<0.0005	0.13	
Count	4	4	4	4	4	4	4	4	
High	4978.00	74.3	6.8	61.0	0.100	0.108	< 0.0005	0.19	
Low	1489.00	35.1	6.6	26.0	0.022	0.028	< 0.0005	0.13	
Mean	2531.25	55.6	6.7	44.8	0.052	0.055	< 0.0005	0.17	
High Limit Low Limit			8.5 6.5	218	1.000	1.000	0.0025	0.49	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Mn	U
	mg/L	mg/L
2017-02	0.082	0.0023
2017-05	0.079	0.0026
2017-08	0.050	0.0024
2017-11	0.036	0.0015
Count	4	4
High	0.082	0.0026
Low	0.036	0.0015
Mean	0.062	0.0022
High Limit	0.800	0.0150
Lim Ex	0.000	0.0130
Frequency	0%	0%
10x Lim Ex	0	0
Frequency	0%	0%

Q-20 Evans Lake Outlet to Dunlop Lake

Month	FLOW	hard	pHF	SO4	Ra	Ва	Со	Fe	
	L/s	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	
2017-11	45.00	37.1	6.9	19.0	<0.007	0.018	<0.0005	0.04	
Count	1	1	1	1	1	1	1	1	
High	45.00	37.1	6.9	19.0	< 0.007	0.018	< 0.0005	0.04	
Low	45.00	37.1	6.9	19.0	< 0.007	0.018	< 0.0005	0.04	
Mean	45.00	37.1	6.9	19.0	<0.007	0.018	<0.0005	0.04	
High Limit Low Limit			8.5 6.5	128.0	1.000	1.000	0.0025	0.49	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Mn	U
	mg/L	mg/L
2017-11	0.030	<0.0005
Count	1	1
High	0.030	< 0.0005
Low	0.030	< 0.0005
Mean	0.030	< 0.0005
High Limit	0.800	0.0150
Lim Ex	0	0
Frequency	0%	0%
10x Lim Ex	0	0
Frequency	0%	0%

SC-01 Westner Lake Outlet

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-11	26.1	6.9	16.0	<0.007	0.009	<0.0005	0.07	0.010	
Count	1	1	1	1	1	1	1	1	
High	26.1	6.9	16.0	< 0.007	0.009	< 0.0005	0.07	0.010	
Low	26.1	6.9	16.0	< 0.007	0.009	< 0.0005	0.07	0.010	
Mean	26.1	6.9	16.0	<0.007	0.009	<0.0005	0.07	0.010	
High Limit Low Limit		8.5 5.2	218	1.000	1.000	0.0025	1.69	0.800	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-11	<0.0005
Count	1
High	< 0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-01 Quirke Lake Outlet

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-10	38.3	6.9	31.0	0.028	0.035	<0.0005	<0.02	0.003	
Count	1	1	1	1	1	1	1	1	
High	38.3	6.9	31.0	0.028	0.035	< 0.0005	< 0.02	0.003	
Low	38.3	6.9	31.0	0.028	0.035	< 0.0005	< 0.02	0.003	
Mean	38.3	6.9	31.0	0.028	0.035	<0.0005	<0.02	0.003	
High Limit Low Limit		8.5 6.5	218	1.000	1.000	0.0025	0.49	0.800	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-10	0.0011
Count	1
High	0.0011
Low	0.0011
Mean	0.0011
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-03

Month	hard	pHF	SO4	TOXCD	TOXDM	Ra	Ва	Co	
	mg/L		mg/L	IC25	%	Bq/L	mg/L	mg/L	
2017-12	37.1	6.9	24.0	100	0	0.012	0.036	<0.0005	
Count	1	1	1	1	1	1	1	1	
High	37.1	6.9	24.0	100	0	0.012	0.036	< 0.0005	
Low	37.1	6.9	24.0	100	0	0.012	0.036	< 0.0005	
Mean	37.1	6.9	24.0	100	0	0.012	0.036	< 0.0005	
High Limit Low Limit		8.5 6.5	128.0			1.000	1.000	0.0025	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Fe	Mn	U	TOXRT
	mg/L	mg/L	mg/L	%
2017-12	0.03	0.011	0.0009	0
Count	1	1	1	1
High	0.03	0.011	0.0009	0
Low	0.03	0.011	0.0009	0
Mean	0.03	0.011	0.0009	0
High Limit	0.49	0.800	0.0150	
Lim Ex	0	0	0	0
Frequency	0%	0%	0%	0%
10x Lim Ex	0	0	0	0
Frequency	0%	0%	0%	0%

SR-06 McCabe Lake Outlet

Month	FLOW	hard	pHF	SO4	TOXCD	TOXDM	TOXRT	Ra	
	L/s	mg/L		mg/L	IC25	%	%	Bq/L	
2017-05	406.20	60.4	7.1	35.0				0.066	
2017-10	785.00	48.0	6.9	35.0				0.099	
2017-11	1056.00	52.0	7.1	32.0				0.092	
2017-12	1121.00	49.4	7.0	32.0	100	0	0	0.099	
Count	4	4	4	4	1	1	1	4	
High	1121.00	60.4	7.1	35.0	100	0	0	0.099	
Low	406.20	48.0	6.9	32.0	100	0	0	0.066	
Mean	842.05	52.5	7.0	33.5	100	0	0	0.089	
High Limit Low limit			8.5 6.5	309				1.000	
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	Ва	Co	Fe	Mn	U
	mg/L	mg/L	mg/L	mg/L	mg/L
2017-05	0.483	<0.0005	0.03	0.009	0.0007
2017-10	0.633	< 0.0005	< 0.02	0.016	0.0007
2017-11	0.631	< 0.0005	0.03	0.011	
2017-12	0.676	<0.0005	0.04	0.007	0.0007
Count	4	4	4	4	3
High	0.676	< 0.0005	0.04	0.016	0.0007
Low	0.483	< 0.0005	< 0.02	0.007	0.0007
Mean	0.606	< 0.0005	0.03	0.011	0.0007
High Limit	1.000	0.0025	0.49	0.800	0.0150
Lim Ex	0	0	0	0	0
Frequency	0%	0%	0%	0%	0%
10x Lim Ex	0	0	0	0	0
Frequency	0%	0%	0%	0%	0%

SR-08 Nordic Lake Outlet

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-02	180.0	6.8	150.0	0.023	0.018	<0.0005	0.04	0.053	
2017-05	179.0	7.2	140.0	0.022	0.015	< 0.0005	0.03	0.019	
2017-08	195.0	7.5	160.0	0.039	0.017	< 0.0005	0.03	0.013	
2017-11	191.0	7.0	150.0	0.020	0.018	<0.0005	0.09	0.058	
Count	4	4	4	4	4	4	4	4	
High	195.0	7.5	160.0	0.039	0.018	< 0.0005	0.09	0.058	
Low	179.0	6.8	140.0	0.020	0.015	< 0.0005	0.03	0.013	
Mean	186.3	7.1	150.0	0.026	0.017	<0.0005	0.05	0.036	
High Limit		8.5	429	1.000	1.000	0.0025	0.49	0.800	
Low Limit		6.5							
Lim Ex	0	0	4	0	0	0	0	0	
Frequency	0%	0%	100%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-02	0.0009
2017-05	0.0009
2017-08	0.0011
2017-11	0.0009
_	
Count	4
High	0.0011
Low	0.0009
Mean	0.0009
112.1.12.2	0.0450
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-15 May Lake

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	57.3	7.1	31.0	0.056	0.133	<0.0005	<0.02	0.006	
2017-10	47.3	6.8	33.0	0.082	0.164	<0.0005	< 0.02	0.003	
Count	2	2	2	2	2	2	2	2	
High	57.3	7.1	33.0	0.082	0.164	< 0.0005	< 0.02	0.006	
Low	47.3	6.8	31.0	0.056	0.133	< 0.0005	< 0.02	0.003	
Mean	52.3	6.9	32.0	0.069	0.149	<0.0005	< 0.02	0.005	
High Limit Low Limit		8.5 6.5	128.0	1.000	1.000	0.0025	0.49	0.800	
Lim Ex	0	0.5	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	<0.0005
2017-10	<0.0005
0	0
Count	2
High	<0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-16 Fox Creek @ Hwy 108

Month hard		pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-02	9.9	5.7	1.6	<0.007	0.008	0.0005	1.19	0.047	
2017-05	6.1	5.8	1.0	< 0.007	< 0.005	< 0.0005	0.38	0.017	
2017-08	8.7	5.8	0.4	< 0.007	0.009	0.0011	1.89	0.068	
2017-11	4.9	5.3	1.2	<0.007	<0.005	<0.0005	0.31	0.021	
Count	4	4	4	4	4	4	4	4	
High	9.9	5.8	1.6	< 0.007	0.009	0.0011	1.89	0.068	
Low	4.9	5.3	0.4	< 0.007	< 0.005	< 0.0005	0.31	0.017	
Mean	7.4	5.7	1.1	<0.007	0.007	0.0006	0.94	0.038	
High Limit		8.5	128.0	1.000	1.000	0.0025	1.69	0.800	
Low Limit		5.2							
Lim Ex	0	0	0	0	0	0	1	0	
Frequency	0%	0%	0%	0%	0%	0%	50%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-02	<0.0005
2017-05	< 0.0005
2017-08	< 0.0005
2017-11	< 0.0005
_	
Count	4
High	<0.0005
Low	<0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-17 Unnamed Creek Drain Lake 3 @ Hwy 108

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-02	17.2	5.7	4.3	< 0.007	0.035	0.0009	0.51	0.053	
2017-05	11.2	6.2	2.8	0.008	0.021	0.0006	0.51	0.034	
2017-08	11.9	5.9	1.2	< 0.007	0.021	0.0010	1.52	0.069	
2017-11	7.0	5.4	3.0	< 0.007	0.012	0.0006	0.37	0.035	
Count	4	4	4	4	4	4	4	4	
High	17.2	6.2	4.3	0.008	0.035	0.0010	1.52	0.069	
Low	7.0	5.4	1.2	< 0.007	0.012	0.0006	0.37	0.034	
Mean	11.8	5.8	2.8	0.007	0.022	0.0008	0.73	0.048	
High Limit		6.5	128.0	1.000	1.000	0.0025	1.69	0.800	
Low Limit		5.2							
Lim Ex	0	0	0	0	0	0	3	0	
Frequency	0%	0%	0%	0%	0%	0%	75%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-02	<0.0005
2017-05	< 0.0005
2017-08	< 0.0005
2017-11	< 0.0005
_	
Count	4
High	<0.0005
Low	<0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-18 Jim Christ Lake Outlet

Month	hard	pHF	SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-05	10.5	6.8	4.0	< 0.007	0.043	<0.0005	0.03	0.008	
2017-11	10.2	6.7	4.0	<0.007	0.043	<0.0005	0.10	0.042	
Count	2	2	2	2	2	2	2	2	
High	10.5	6.8	4.0	< 0.007	0.043	< 0.0005	0.10	0.042	
Low	10.2	6.7	4.0	< 0.007	0.043	< 0.0005	0.03	0.008	
Mean	10.4	6.8	4.0	<0.007	0.043	<0.0005	0.07	0.025	
High Limit		8.5	128.0	1.000	1.000	0.0025	0.49	0.800	
Low Limit		6.5							
Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-05	<0.0005
2017-11	<0.0005
0	•
Count	2
High	<0.0005
Low	< 0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

SR-19 Inlet to Elliot Lake

Month	th hard		SO4	Ra	Ва	Co	Fe	Mn	
	mg/L		mg/L	Bq/L	mg/L	mg/L	mg/L	mg/L	
2017-02	14.5	7.1	3.5	< 0.007	0.021	<0.0005	0.34	0.026	
2017-05	13.3	7.0	2.8	< 0.007	0.017	< 0.0005	0.26	0.030	
2017-08	16.7	6.9	2.7	0.010	0.021	< 0.0005	0.60	0.040	
2017-11	13.3	6.9	3.0	<0.007	0.017	<0.0005	0.23	0.028	
Count	4	4	4	4	4	4	4	4	
High	16.7	7.1	3.5	0.010	0.021	< 0.0005	0.60	0.040	
Low	13.3	6.9	2.7	< 0.007	0.017	< 0.0005	0.23	0.026	
Mean	14.4	7.0	3.0	0.008	0.019	<0.0005	0.36	0.031	
High Limit		8.5	128.0	1.000	1.000	0.0025	0.49	0.800	
Low Limit		6.5							
Lim Ex	0	0	0	0	0	0	1	0	
Frequency	0%	0%	0%	0%	0%	0%	25%	0%	
10x Lim Ex	0	0	0	0	0	0	0	0	
Frequency	0%	0%	0%	0%	0%	0%	0%	0%	

Month	U
	mg/L
2017-02	<0.0005
2017-05	< 0.0005
2017-08	< 0.0005
2017-11	< 0.0005
_	
Count	4
High	<0.0005
Low	<0.0005
Mean	<0.0005
High Limit	0.0150
Lim Ex	0
Frequency	0%
10x Lim Ex	0
Frequency	0%

Rio Algom Limited and Denison Mines Inc. 2017 Serpent River Watershed Water Quality Monitoring Results Five Year Annual Average Station D-4

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark 3							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.05
2013		6.8	4.3	< 0.005	< 0.0005	0.013	< 0.0005	0.04	0.017	10.4
2014		7.2	4.6	< 0.005	< 0.0005	0.013	< 0.0005	0.03	0.010	10.0
2015		6.9	3.9	< 0.007	< 0.0005	0.013	< 0.0005	0.03	0.017	9.3
2016		6.8	3.8	< 0.008	< 0.0005	0.013	< 0.0005	0.04	0.016	10.8
2017		6.8	3.5	< 0.007	< 0.0005	0.013	< 0.0005	0.04	0.021	9.6

Notes:

Bold indicates exceedance of evaluation criteria value

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

Rio Algom Limited and Denison Mines Inc. 2017 Serpent River Watershed Water Quality Monitoring Results Five Year Annual Average Station SR-18

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.50
2013		7.0	5.0	< 0.005	< 0.0005	0.051	< 0.0005	0.06	0.025	10.8
2014		6.9	4.3	< 0.005	< 0.0005	0.043	< 0.0005	0.07	0.013	10.5
2015		6.9	4.6	< 0.007	< 0.0005	0.048	< 0.0005	0.07	0.030	10.4
2016		7.0	4.5	< 0.008	< 0.0005	0.048	< 0.0005	0.05	0.015	11.5
2017		6.8	4.0	< 0.007	< 0.0005	0.043	< 0.0005	0.07	0.025	10.4

Notes:

Bold indicates exceedance of evaluation criteria value

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 202016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardnes s (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ² Lake benchmark ³	5.2						1.69 0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.50
2013		7.1	4.2	< 0.005	< 0.0005	0.022	< 0.0005	0.21	0.028	16.2
2014		7.0	3.7	< 0.005	< 0.0005	0.022	< 0.0005	0.31	0.038	15.0
2015		7.1	4.0	< 0.007	< 0.0005	0.025	< 0.0005	0.40	0.050	18.2
2016		6.8	4.0	< 0.008	< 0.0005	0.026	< 0.0005	0.35	0.054	16.0
2017		7.0	3.0	0.008	< 0.0005	0.019	< 0.0005	0.36	0.031	14.4

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria ¹	Wetland and lake benchmarks	6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark 3							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		5.4	1.4	< 0.005	< 0.0005	0.007	0.0008	1.11	0.057	8.4
2014		5.5	1.1	< 0.005	< 0.0005	0.007	0.0007	0.77	0.045	6.9
2015		5.6	1.1	< 0.007	< 0.0005	0.006	0.0007	0.97	0.044	7.2
2016		5.8	1.3	< 0.008	< 0.0005	0.007	0.0006	0.67	0.032	8.0
2017		5.7	1.1	< 0.007	< 0.0005	0.007	0.0007	0.94	0.038	7.4

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland/stream stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		5.6	2.8	< 0.005	< 0.0005	0.018	0.0010	1.54	0.053	10.5
2014		5.4	2.8	< 0.005	< 0.0005	0.018	0.0011	1.19	0.067	10.2
2015		5.6	2.6	< 0.007	< 0.0005	0.017	0.0010	1.11	0.061	10.2
2016		5.8	2.5	0.009	< 0.0005	0.022	0.0010	1.32	0.064	12.8
2017		5.8	2.8	0.007	< 0.0005	0.022	0.0008	0.73	0.048	11.8

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland/stream stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks		6.5	128.0	1.000	0.0150	1.000	0.0025		0.800	
	Wetland benchmark ²		5.2						1.69		
	Lake benchmark ³								0.49		
MDL ⁴			0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		6561.5	6.9	13.9	0.037	0.0017	0.040	< 0.0005			22.5
2014		5672.0	7.1	14.9	0.053	0.0017	0.059	< 0.0005			
2015		2144.3	6.9	13.8	0.057	0.0015	0.068	< 0.0005			
2016		1884.0	6.8	14.1	0.069	0.0015	0.047	< 0.0005	0.08	0.047	26.6
2017		4843.0	6.8	11.3	0.040	0.0013	0.045	< 0.0005	0.07	0.026	20.5

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

Bold indicates exceedance of evaluation criteria value

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks		6.5	309.0	1.000	0.0150	1.000	0.0025	•	0.800	-
	Wetland benchmark ² Lake benchmark ³		5.2						1.69 0.49		
MDL ⁴	Lake benefitian		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		317.3	6.6	17.5	0.006	< 0.0005	0.013	< 0.0005	0.22	0.091	25.8
2014		396.4	6.6	33.0	0.006	< 0.0005	0.017	0.0006	0.34	0.215	42.2
2015		167.9	6.7	22.8	0.007	< 0.0005	0.017	0.0007	0.45	0.267	53.3
2016		95.3	6.6	88.8	0.011	< 0.0005	0.022	0.0013	0.54	0.458	100.9
2017		151.9	6.7	18.8	< 0.007	< 0.0005	0.013	< 0.0005	0.19	0.102	28.7

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland/stream stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵	Ra	U	Ba	Co	Fe	Mn ⁵	Hardness
A			(mg/L)	(Bq/L)	(mg/L)	(mg/L	(mg/L)	(mg/L)	(mg/L)	(mg/)
Assessment Criteria ¹	Wetland and lake benchmarks	6.5	309.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005		0.02	0.002	0.5
2012		6.8	56.5	0.009	< 0.0005	0.018	< 0.0005	0.22	0.131	33.7
2013		6.6	17.0	< 0.005	< 0.0005	0.013	< 0.0005	0.23	0.091	23.9
2014		6.7	15.0	< 0.005	< 0.0005	0.012	< 0.0005	0.16	0.072	21.3
2015		6.8	21.0	< 0.007	< 0.0005	0.012	< 0.0005	0.13	0.066	26.1
2016		6.6	54.0	0.010	< 0.0005	0.017	0.0006	0.28	0.160	64.6

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness mg/L
Assessment Criteria ¹	Wetland and lake benchmarks	6.5	309.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2012		5.7	< 0.1	< 0.005	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002	< 0.5
2013		6.2	< 0.1	< 0.005	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002	< 0.5
2014		5.4	< 0.1	< 0.005	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002	< 0.5
2015		5.4	0.3	< 0.007	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002	< 0.5
2016		5.7	< 0.1	< 0.008	< 0.0005	< 0.005	< 0.0005	< 0.02	< 0.002	< 0.5

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria ¹	Wetland and lake benchmarks		6.5	309.0	1.000	0.0150	1.000	0.0025		0.80	-
	Wetland benchmark ²		5.2						1.69		
	Lake benchmark ³								0.49		
MDL ⁴			0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		309.9	7.2	97.5	0.103	0.0012	0.016	< 0.0005	0.33		119.0
2014		286.3	6.8	73.5	0.101	0.0013	0.016	0.0007	1.68		141.8
2015		126.9	7.0	94.8	0.135	0.0007	0.020	< 0.0005	0.39		93.0
2016		118.5	7.0	58.8	0.131	0.0006	0.018	< 0.0005	0.34	0.020	80.6
2017		338.7	6.8	59.8	0.193	0.0008	0.017	< 0.0005	0.60	0.037	83.5

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	218.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark 3							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		6.8	13.0	0.022	0.0035	0.017	< 0.0005	0.56		36.5
2014		6.9	10.3	0.014	0.0032	0.015	< 0.0005	0.82		
2015		6.8	11.1	0.013	0.0025	0.015	< 0.0005	0.50		
2016		6.7	11.4	0.021	0.0026	0.017	< 0.0005	0.43	0.017	44.4
2017		6.8	10.0	0.016	0.0034	0.015	< 0.0005	0.58	0.070	36.3

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria ¹	Wetland and lake benchmarks	6.5	218.0	1.000	0.0050	1.000	0.0025		0.8	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2012		6.7	16.5	0.022	0.0042	0.017	0.0007	0.66		40.5
2013		7.0	11.5	0.022	0.0027	0.015	< 0.0005	0.40		32.8
2014		7.3	10.1	0.015	0.0033	0.016	< 0.0005	0.32		31.1
2015		6.9	11.0	0.014	0.0026	0.014	< 0.0005	0.33		
2016		6.7	13.0	0.026	0.0021	0.018	< 0.0005	0.41	0.136	38.8

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

²Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF		SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	(Ba (mg/L)	Co (mg/L)	(1	Fe mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5		218.0	1.000	0.0150		1.000	0.0025			0.800	-
	Wetland benchmark ²	5.2									1.69		
	Lake benchmark 3										0.49		
MDL ⁴		0.1		0.1	0.005	0.0005		0.005	0.0005		0.02	0.002	0.5
2012		5.8	<	0.1	< 0.005	< 0.0005	<	0.005	< 0.0005	<	0.02		< 0.5
2013		5.6	<	0.1	< 0.005	< 0.0005	<	0.005	< 0.0005	<	0.02		< 0.5
2014		5.6	<	0.1	< 0.005	< 0.0005	<	0.005	< 0.0005	<	0.02		
2015		5.6	<	0.1	< 0.007	< 0.0005	<	0.005	< 0.0005	<	0.02		
2016		5.3	<	0.1	< 0.008	< 0.0005	<	0.005	< 0.0005	<	0.02	< 0.002	< 0.5

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks		6.5	218.0	1.000	0.0150	1.000			0.800	-
	Wetland benchmark 2		5.2						1.69		
	Lake benchmark ³								0.49		
MDL 4			0.1	0.1	0.005	0.0005	0.005		0.02	0.002	0.5
2013		6687.50	6.8	37.0	0.041	0.0025	0.044	< 0.0005			45.1
2014		5819.50	7.0	37.0	0.068	0.0021	0.073	< 0.0005			47.4
2015		2187.75	6.8	47.8	0.069	0.0027	0.095	< 0.0005			59.0
2016		1956.25	6.6	82.3	0.077	0.0027	0.097	< 0.0005	0.08	0.034	92.8
2017		2531.30	6.7	44.8	0.052	0.0015	0.055	< 0.0005	0.17	0.036	55.6

Notes:

Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent.

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks		6.5	218.0	1.000	0.0150	1.000			0.800	
	Wetland benchmark ²		5.2						1.69		
	Lake benchmark ³								0.49		
MDL ⁴			0.1	0.1	0.005	0.0005	0.005		0.02	0.002	
2013		15.0	7.2	21.0	< 0.005	< 0.0005	0.021	< 0.0005			39.5
2014		57.3	7.0	21.0	< 0.005	< 0.0005	0.018	< 0.0005			
2015		4.0	7.0	21.0	< 0.008	< 0.0005	0.018	< 0.0005			
2016		2.0	6.8	22.0	< 0.008	< 0.0005	0.020	< 0.0005	< 0.02	0.014	40.0
2017		45.0	6.9	19.0	< 0.007	< 0.0005	0.018	< 0.0005	0.04	0.030	37.1

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	(1	Ra Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn ⁵ (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	218.0		1.000	0.0150	1.000	0.0025		0.800	
	Wetland benchmark ²	5.2							1.69		
	Lake benchmark ³								0.49		
MDL ⁴		0.1	0.1	(0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		7.1	23.0	< (0.005	< 0.0005	0.012	< 0.0005	0.12		33.8
2014		7.0	21.0	(0.006	< 0.0005	0.010	< 0.0005	0.12		
2015		7.0	21.0	< (800.0	< 0.0005	0.010	< 0.0005	0.07		
2016		6.9	20.0	< (0.008	< 0.0005	0.010	< 0.0005	0.06	0.008	31.0
2017		6.9	16.0	< (0.007	< 0.0005	0.009	< 0.0005	0.07	0.010	26.1

Notes

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent.

Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	218.0	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ² Lake benchmark ³	5.2						1.69 0.49		
MDL ⁴		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		6.9	39.0	0.025	0.0012	0.038	< 0.0005			46.3
2014		6.9	34.0	0.017	0.0013	0.038	< 0.0005			
2015		6.9	36.0	0.019	0.0014	0.039	< 0.0005			
2016		6.8	33.0	0.026	0.0013	0.036	< 0.0005	< 0.02	0.003	40.0
2017		6.9	31.0	0.028	0.0011	0.035	< 0.0005	<0.02	0.003	38.3

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		FLOW (L/s)	pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn (mg/L	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks		6.5	309.0	1.000	0.0150	1.000	0.0025		0.8	
	Wetland benchmark ²		5.2						1.69		
	Lake benchmark ³								0.49		
MDL ⁴			0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		1815.1	7.2	58.5	0.052	0.0010	0.274	< 0.0005			69.8
2014		953.9	6.8	47.5	0.057	0.0009	0.313	< 0.0005			
2015		276.0	7.2	46.5	0.064	0.0008	0.450	< 0.0005			
2016		494.5	6.9	39.3	0.074	0.0007	0.512	< 0.0005	0.03	0.014	53.2
2017		842.1	7.0	35.5	0.089	0.0007	0.606	< 0.0005	0.03	0.011	52.6

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	429	1.000	0.0150	1.000	0.0025		0.800	-
	Wetland benchmark ²	5.2						1.69		
	Lake benchmark ³							0.49		
MDL ⁴		0.1		0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2013		7.2	171	0.025	0.0010	0.020	< 0.0005			185.3
2014		7.0	165	0.026	0.0009	0.018	< 0.0005			176.3
2015		7.1	155	0.028	0.0009	0.018	< 0.0005			170.0
2016		6.8	150	0.029	0.0009	0.017	< 0.0005	0.05	0.033	178.5
2017		7.1	150	0.026	0.0009	0.017	< 0.0005	0.05	0.036	186.3

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

⁴ Method Detection Limits as per Table 5.2 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

⁵ Sulphate and Manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

YEAR		pHF	SO4 ⁵ (mg/L)	Ra (Bq/L)	U (mg/L)	Ba (mg/L)	Co (mg/L)	Fe (mg/L)	Mn (mg/L)	Hardness (mg/L)
Assessment Criteria 1	Wetland and lake benchmarks	6.5	128.0	1.000	0.0050	1.000	0.0025		0.8	
	Wetland benchmark ² Lake benchmark ³	5.2						1.69 0.49		
MDL 4		0.1	0.1	0.005	0.0005	0.005	0.0005	0.02	0.002	0.5
2016		7.0	36.5	0.049	< 0.0005	0.139	< 0.0005	< 0.02	0.005	51.8
2017		6.9	32	0.069	<0.0005	0.149	<0.0005	< 0.02	0.005	52.3

Notes:

¹ Assessment criteria as per Table 4.5 Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow, 2016)

² Benchmark applies to wetland stations: M-01, DS-18, SC-01.

³ Benchmark applies to lake stations: D-5, D-6, Q-09, Q-20, SR-01, SR-06, SR-08.

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⁵ Sulphate and manganese criteria taken from Table B.1, Appendix B, Cycle 4 Study Design for the SRWMP, SAMP and TOMP (Minnow,2016). Parameters are hardness dependent. Variation in number of significant figures reflect MDL's at the time of reporting. In 2006, laboratory reported MDL's were standardized to achieve consistency and meet program requirements, as per Cycle 2 Interpretive Report (Minnow 2005).

APPENDIX VI Interim Public Dose Estimation for the Closed Mines of the Serpent River Watershed



INTERIM PUBLIC DOSE ESTIMATION FOR THE CLOSED MINES OF THE SERPENT RIVER WATERSHED

Report prepared for:

Rio Algom Limited PO Box 38 Elliot Lake, Ontario P5A 2J6

and

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16-2226 February 2018



INTERIM PUBLIC DOSE ESTIMATION FOR THE CLOSED MINES OF THE SERPENT RIVER WATERSHED

Janeen Tang, B.A.Sc.,M.E.S., P.Eng.
Project Manager and Principal Investigator

Janeen Tang

Don Hart, Ph.D.

Project Principal and Reviewer



EXECUTIVE SUMMARY

The Canadian Nuclear Safety Commission (CNSC) has asked Rio Algom Limited and Denison Mines Inc. to undertake annual reporting of radiation dose to the public associated with their closed uranium mine sites in the Serpent River Watershed. The annual dose reporting will be based on periodic updates undertaken as part of the five-year State of the Environment (SOE) reports.

This interim public dose estimation is intended to provide interim dose values for the period 2017 – 2020 when the next SOE is scheduled for completion. The intention is to estimate realistic doses for a representative person residing in the City of Elliot Lake to be included in annual Serpent River Watershed Monitoring reports effective 2017. Elliot Lake is the only lake in the watershed with an urban community. The residents of the City are potentially exposed to radioactive substances via both Elliot Lake water and recreational use of mine properties, and are considered to be the population with the greatest potential for exposure to radiation and radioactive materials from the closed mine sites.

Ingestion of drinking water from Elliot Lake, and ingestion of fish caught in this and other lakes downstream of the Tailings Management Areas (TMAs) were identified as key ingestion pathways based on upper bound public dose estimates prepared for SOE reports for the Serpent River Watershed. Radon and direct gamma pathways were identified as key pathways based on upper bound dose estimates for people walking near TMAs prepared by the CNSC.

Site-specific surveys of residents were undertaken by Rio Algom in 2016 to characterize resident exposure pathways and habits relevant to exposure, and monitoring was undertaken to characterize mine site gamma and radon fields, as well as drinking water radionuclide concentrations, to update the characterization of the levels of public exposure. This report includes the results of the 2016 site-specific surveys and monitoring programs, as well as an interim public dose estimation, based on current understanding of human receptors and key exposure pathways.

The interim monitoring program to support public dose estimation for a representative Elliot Lake resident included:

- Surveys of City of Elliot Lake residents to refine estimates of time spent on roads and trails near the TMAs and estimates of fish consumption from different lakes in the Serpent River watershed;
- Monitoring of radon and gamma on roads and trails near TMAs often used by walkers and hikers; and
- Monitoring of appropriate U-238 series radionuclides in drinking water from the City of Elliot Lake Water Treatment Plant, after treatment.

i



The 2016 data from these surveys and monitoring programs were used in the interim public dose estimation. Available data for radionuclides in sport fish from 2005 were also used. The sport fish data pertain to the lakes most used for fishing but should be updated.

Based on the interim public dose calculations, it may be concluded that:

- Public dose to the representative person is approximately 0.012 mSv/a, after correction for background exposure.
- This value is based on available measurements of radon and direct gamma near TMAs, and U-238 series radionuclides in treated drinking water and sport fish, as well as survey information and several assumptions for exposure factors.

The public dose estimation will next be updated as part of the 2020 State of the Environment report. Changes to the monitoring data are projected to include updated sport fish tissue analysis.



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1.0 INTRODUCTION

1.1 Background

The Canadian Nuclear Safety Commission (CNSC) has asked Rio Algom Limited and Denison Mines Inc. to undertake annual reporting of radiation dose to the public associated with their closed uranium mine sites in the Serpent River Watershed.

State of the Environment (SOE) reports for the watershed have focused on demonstrating upper bounds of public dose, using rather conservative assumptions for hypothetical human residents on lakes downstream of the Tailings Management Areas (TMAs). The CNSC (2002) also estimated upper bound doses from recreational use of TMAs in and around Elliot Lake, based on conservative use assumptions. Neither of these dose estimates are considered to be realistic estimates of public dose.

The intention of this report is to present a more realistic public dose for a representative person exposed to radioactivity from the closed mine properties. This will be included in annual Serpent River Watershed Monitoring reports effective 2017. The value reported in the 2017 annual report will be interim, based on the radon and gamma survey data along with two rounds of drinking water quality data collected in 2016. In the 2017, 2018 and 2019 annual reports, the interim public dose value from 2017 will continue to be reported. Updates to the public dose estimation will be included in the next State of the Environment report whose field program is scheduled for the fall of 2019.

Elliot Lake is the only lake in the watershed with an urban community. The residents of the city are potentially exposed to radioactive substances via consumption of Elliot Lake water and recreational use of mine properties, and are considered to be the population with the greatest potential for exposure to radiation and radioactive materials from the closed mine sites.

A preliminary design for a monitoring program to support public dose estimation was prepared in early 2016 (EcoMetrix, 2016). Based on this plan, site-specific surveys of residents were completed in 2016 to characterize their exposure pathways and habits relevant to exposure. Additional radiological monitoring in 2016 included radon and gamma monitoring on roads and trails near TMAs often used by walkers and hikers and radionuclide monitoring of drinking water. This information has been used in conjunction with historic fish tissue data to estimate an interim public dose for a representative resident of the City of Elliot Lake.

1.2 Objectives

The objective of this report is to document interim survey and monitoring data relevant to public exposure, and use it to derive an interim realistic public dose estimate for a representative person in the population group living near the closed mine properties in the



Serpent River watershed. It is recognized that some revisions to the estimated public dose may be appropriate once additional data become available, and these will be addressed as part of the updates to the representative public dose in the 2020 State of the Environment Report.



2.0 THE REPRESENTATIVE PERSON AND MAIN EXPOSURE PATHWAYS

2.1 The Representative Person

In estimating public dose for comparison to a dose constraint, the dose is estimated for a "representative person" with characteristics that reflect those of the group that receives the highest dose (ICRP, 2007). The representative person is equivalent to, and replaces, the "average member of the critical group" which was previously used for determining compliance with a public dose constraint (ICRP, 1986).

The critical group within which the representative person is defined must be large enough to support reliable characterization of typical habits relevant to exposure, and should be relatively homogeneous. ICRP (1986) has defined homogeneity to mean that the individual doses within the group lie substantially within a range of a factor of ten, provided that the mean is less than one-tenth of the dose limit. The preliminary survey of 300 Elliot Lake residents indicates that, of those using TMAs for walking and hiking, the use rates vary within a factor of ten. Since this is the dominant exposure pathway of dose, the Elliot Lake group was considered to be homogeneous.

The group size and homogeneity conditions imply that the representative person should not be characterized based on single individuals or households with extreme behaviors. Rather, the representative person can reflect an average across distinct practices within the group. Site-specific surveys on habits relevant to exposure should be conducted to characterize the representative person. Surveys should address the use of local food and water resources, as appropriate. The Elliot Lake survey addressed local fish consumption, as well as TMA use for walking and hiking.

In characterizing the representative person, averaging should not occur across age classes. The ICRP (2007) considers three age classes, for which intake rates and dose coefficients have been defined. Nominal ages are 1 year (age 0 to 5), 10 years (age 6-15) and adult (age 16 to 70). These age classes may be designated infant, child and adult (CSA, 2014). Since Elliot Lake is a retirement community, the adult age class is dominant, with only 10% of the population in the 0-14 age group according to the 2011 census. The preliminary survey of 300 residents indicated that only 6 out of 300 respondents had children under the age of 16 who walk or hike around the closed mine properties, and only 7 had children who eat locally caught fish. The small sample sizes make it difficult to reliably characterize TMA or local fish use rates for child or infant age groups. Therefore, only adult doses were estimated.

2.2 Main Exposure Pathways

Upper bound ingestion doses to hypothetical human residents on lakes downstream of the TMAs were estimated by EcoMetrix (2011) and included in Appendix F of the 2011 SOE



report (Minnow, 2011). While these estimates are conservative, and are not considered suitable as public dose estimates, they provide a preliminary indication of key ingestion pathways. The dose estimate for a hypothetical resident on Elliot Lake, after correction for background, was 0.0244 mSv/a. The ingestion dose estimates were based on 1.5 L/day of lake water intake, 2.92 kg/year of local fish consumption, 1 kg/year of local waterfowl (mallard) consumption and 1 kg/year of local moose meat consumption. Of the pathways considered, intake of drinking water and fish accounted for almost 99% of the incremental dose (73% from drinking water, 26% from fish). These two main exposure pathways have been included in the interim public dose calculations.

Upper bound doses arising from exposure to radon and gamma radiation while walking near TMAs were estimated by the CNSC (2002) as 0.04 and 0.06 mSv/a (incremental), respectively, based on an assumed 200 hours each year at the location of highest measured radon and gamma radiation (Lacnor and Nordic TMAs). The gamma dose estimate makes no allowance for the cover that was placed on the tailings after the gamma survey. The incremental dose from radon was estimated as 0.016 mSv/a for a person at Nordic Lake. The radon dose estimates assume full progeny ingrowth. While these estimates are overly conservative, and not suitable as public dose estimates, they suggest that radon and direct gamma pathways should be included in the public dose calculations.

The assumed water ingestion rate of 1.5 L/day (Health Canada, 1995) is an average value for adults in the general population. The water supply for Elliot Lake residents comes from Elliot Lake. Water consumption rates are physiologically driven, thus it is reasonable to apply this rate to the Elliot Lake water supply for Elliot Lake residents.

The assumed fish ingestion rate of 2.92 kg/year (8 g/day, U.S. EPA, 1997) is a value for freshwater anglers. The U.S. EPA (2011) also cites a value of 5 g/day for anglers around Lake Ontario. The assumption that all fish are taken from Elliot Lake is probably overly conservative. While Elliot Lake and other local lakes may be used, it is likely that much of the fish consumption is not of local origin. The 2016 site-specific survey has been used in the present assessment to clarify local amounts of fish consumption.

The concentrations of radionuclides in water that have been used in historical dose calculations are either measured values in Elliot Lake water, or values estimated from sediments and partition coefficients. The water quality monitoring data are often reported as "non-detects", which are values below the reporting limit. The use of water quality monitoring data for Elliot Lake means that there has been no accounting for water treatment. 2016 treated water monitoring results have been used in the present assessment for the calculation of dose from municipal drinking water.

The concentrations of radionuclides in sport fish that have been used in the dose calculations are either measured values in sport fish from Elliot Lake, or values estimated from water and bioaccumulation factors. The fish chemistry data are often reported as non-detects. Only Unat and Ra-226 have actually been measured in sport fish. Some of the bioaccumulation factors used for sport fish were estimated from forage fish values.



Measured activities in sport fish should be considered for the calculation of dose from fish ingestion.



3.0 SURVEYS AND MONITORING TO CHARACTERIZE EXPOSURE

The site-specific surveys and monitoring programs support the estimation of realistic public dose to a representative person residing in the City of Elliot Lake. Based on the main exposure pathways identified in Section 2.2, the surveys and monitoring program included the following components: surveys to refine estimates of time spent hiking at TMAs, surveys to refine estimates of fish consumed from lakes downstream of TMAs, monitoring of radon and gamma where people walk, hike or otherwise use trails and roads at TMAs, and monitoring of radionuclides in drinking water. Historic data have been used for sport fish tissue concentrations, though measured activities in sport fish will be collected as part of the 2019 biological component of the Cycle 5 State of the Environment study and a revised public dose included in the 2020 Cycle 5 State of the Environment report. The design of surveys and monitoring programs is discussed in the following sections.

3.1 Survey of Habits Relevant to Exposure

The intent of the site-specific surveys of residents is to characterize exposure pathways and habits relevant to exposure. The survey form for trail users and fishers was designed to address the following questions:

- How many hours do residents use trails and roads at the TMAs?
- How do people divide their trail use time among the TMAs?
- Where do people fish and how much fish do they consume from each lake?
- Which fish species are consumed?

The answers provided in the survey informed the estimation of public dose and the design of the monitoring programs for radon, gamma and sport fish. The survey was administered to residents of the Elliot Lake area as part of a larger community survey conducted on behalf of Rio Algom Limited by Globescan. A screening question was included at the beginning of the survey to determine whether the respondent is a resident of Elliot Lake. Data were collected from one respondent per household who responded on behalf of the entire household.

The survey questions and results are provided in Appendix A.

It is expected that most people in Elliot Lake will be on municipal water, which is drawn from Elliot Lake and treated prior to distribution. There are some homeowners and cottage owners on the lakefront, who take water directly from the lake, and Quirke Lake is the main area for development. There may also be some people who drink bottled water. In characterizing the representative Elliot Lake resident, it would be reasonable to assume



that this person drinks water from the municipal system. Surveys to investigate the frequency of use of drinking water sources other than the municipal water supply could be considered at a later date, based on information about these sources.

3.2 Measurements of Radon and Direct Gamma

The intent of the radon and direct gamma monitoring program is to characterize levels of exposure to radon and direct gamma for trail users at the TMAs.

Denison Environmental Services (DES, 2016a,b) conducted an initial monitoring program in December 2015 to measure radon and direct gamma radiation during walking surveys at the Lacnor, Milliken, Stanleigh, Quirke, Panel, Nordic, Pronto, Denison and Stanrock TMAs. In 2016, the surveys were repeated quarterly and extended to include the Buckles tailings and the Spanish American TMA. DES provided the data. The surveys were conducted quarterly to assess seasonal variability. Figures showing the location of the walking surveys are provided in **Appendix B**. The Esten Lake boat launch trail was surveyed to estimate background radiation in areas uninfluenced by TMAs. Radon decay products were collected on filter paper with an air sampling pump and then alpha radiation was measured using a scintillometer. Gamma radiation was measured using an SEI Inspector USB multiradiation detector.

Radon was typically highest in October or December and lowest in April or July, with a maximum/minimum ratio ranging between 8 and 28. The gamma field was typically highest in July or October, but was much less variable, and was the dominant component of exposure for a walker near any TMA. The maximum/minimum ratio for the total exposure ranged between 1.1 and 1.9, and averaged 1.4. Since there was little seasonality in total exposure, it was considered acceptable to characterize TMA use on an annual basis.

3.3 Monitoring of Concentrations in Drinking Water and Fish

The intent of the monitoring program for drinking water and sport fish is to characterize levels of exposure to radionuclides through the ingestion pathway.

The water treatment plant for the City of Elliot Lake provides annual reporting on concentrations of uranium in treated water. Recent annual reports indicate that uranium concentrations in municipal drinking water were 0.172 μ g/L on 31 January 2014 and 0.149 μ g/L on 22 July 2015 (City of Elliot Lake, 2014, 2015). Uranium concentrations in municipal drinking water are approximately one tenth of uranium concentrations in untreated water from Elliot Lake, based on 2010 lake water quality used in the previous dose assessment for Elliot Lake (EcoMetrix, 2011).

Health Canada (2009) has suggested that the measured levels of radionuclides in the Elliot Lake water supply likely represent natural background. Although drinking water in Elliot Lake may not be different from background, monitoring of radionuclides in municipal



drinking water was completed in August 2016 and November 2016 to confirm this and provide data to support the estimation of annual public dose.

Detection limits for radionuclides in drinking water were 0.1 μ g/L for U, 0.01 Bq/L for Th-230, 0.005 Bq/L for Ra-226, 0.02 Bq/L for Pb-210, and 0.005 Bq/L for Po-210. Certificates of Analysis for drinking water are provided in **Appendix C**.

Radionuclides in the Th-232 series were not detected in lake waters during the special investigation study (EcoMetrix, 2011). They were elevated in sediments relative to background only in Quirke and May lakes. Using partitioning estimates of concentrations in water, the drinking water dose from the Th-232 series contributed less than 5% of the total ingestion dose for the representative human at Elliot Lake. The analysis of Th-232 radionuclides in drinking water was not considered to be warranted because of its small contribution to total dose.

Concentrations of uranium and Ra-226 were measured in 2004 in northern pike, smallmouth bass and lake trout from lakes downstream of the TMAs (Elliot Lake, Quirke Lake and McCarthy Lake) and in reference lakes (Minnow, 2005).

As part of the 2019 update, the following radionuclides should be analyzed in sport fish: Unat, Th-230, Ra-226, Pb-210 and Po-210. Based on the results of the survey of Elliot Lake residents, fish should be collected from Elliot, Quirke and McCarthy lakes, as these were the lakes in the Serpent River watershed most used by sport fishers. Lake trout and walleye should be targeted as they were identified as the species most consumed by Elliot Lake residents; smallmouth bass and northern pike were next in order of preference, and may also be considered. Radionuclide concentrations in fish are likely to change slowly, following the gradual improvement of water quality in the lakes.

Low detection limits will be needed to obtain quantitative results for radionuclides in fish tissue. Suggested detection limits for fish tissue are: $0.001\mu g/g$ for U, 0.0001 Bq/g for Th-230, 0.0006 Bq/g for Ra-226, 0.001 Bq/g for Pb-210, and 0.0002 Bq/g for Po-210. Prior to collecting fish tissue samples, the analytical laboratory should be consulted regarding detection limits, sample size and other sampling requirements. It may be possible to achieve lower detection limits by increasing the sample volume.

Radionuclides in the Th-232 series were not measured in sport fish during the special investigation study (EcoMetrix, 2011). They were elevated in sediments relative to background only in Quirke and May lakes. Using bioaccumulation estimates of concentrations in sport fish, the sport fish dose from the Th-232 series contributed less than 1% of the total ingestion dose for the representative human at Quirke Lake and 3% of the total ingestion dose for the representative human at May Lake. The analysis of Th-232 radionuclides in sport fish is not considered to be warranted because of its small contribution to total dose.



4.0 INTERIM ESTIMATION OF PUBLIC DOSE

4.1 Overview of Approach

The approach to public dose estimation is intended to capture the main exposure pathways for Elliot Lake residents, as discussed in Section 2.2. Those pathways are exposure to radon and direct gamma radiation while walking near TMAs, ingestion of drinking water from Elliot Lake, and ingestion of fish caught in lakes downstream of the TMAs. The approach is intended to produce a realistic dose estimate, and uses 2016 survey and monitoring data to improve the estimate. An interim dose estimation for adult residents is presented here to illustrate the approach. The adult age class is dominant in Elliot Lake, as noted in Section 2.1; therefore, only adult dose was calculated. The use of survey and monitoring data and the assumptions made in this initial dose estimation are discussed by pathway in the following sections.

4.2 Radon and Direct Gamma Measurements

Measurements of dose from radon progeny in air and from direct gamma radiation while walking on roads and trails near the TMAs were obtained by Denison Environmental Services during four surveys in 2016 (April, July, October, December) (data provided by DES). Radon decay products were collected on filter paper with an air sampling pump and then alpha radiation was measured using a scintillometer. Gamma radiation was measured using an SEI Inspector USB multi-radiation detector. **Table 4-1** summarizes the results.

Measurements at the Esten Lake boat launch trail provide an estimate of background dose while walking in areas uninfluenced by TMAs. The Esten Lake area was chosen because it has similar environmental characteristics to the TMAs, but has no tailings nearby.

A survey of Elliot Lake residents conducted by Rio Algom Limited in 2016 (Appendix A) provides an estimate of the actual number of hours per year spent walking near TMAs for the representative person, and the proportion of that time spent at each TMA. The measured doses recorded for a nominal 200 hours per year at each TMA were adjusted for actual hours, and a weighted average dose across TMAs was calculated using the proportion of time at each TMA. The average number of hours per year walking at TMAs was 110.76 hours (2.13 hours per week) for a typical Elliot Lake resident.

The survey of Elliot Lake residents indicated that Milliken/Sheriff Creek Park was most used for walking and hiking, followed by the Quirke TMA. The use proportions for the TMAs, as reported in the survey, were adjusted up to account for the people who did not know the TMA used. The resulting proportions (45.3% Milliken, 12.6% Quirke, 9.5% Stanleigh, 9.5% Nordic, 8.4% Panel, 7.4% Denison, and 7.4% Stanrock) were used to allot the time spent walking among TMAs, making a weighted average dose from casual access at TMAs for the typical Elliot Lake resident.



Table 4.1: Radon and Direct Gamma Doses from Walking Near TMAs (2016)

Site	Route	Radon Dose (mSv/a*)	Gamma Mean Dose (mSv/a*)	Total Annual Dose (mSv/a*)	Annual Dose for TMA (mSv/a*)
Denison	William's Lake ETP to Settling Pond Berm	0.003818	0.058728	0.062545	
Demison	TMA 1 Treatment Plant to Dam 10	0.001430	0.029773	0.031203	0.04687
Stanrock	Main Gate to Rooster Rock	0.004158	0.043343	0.047501	
Stalliock	Main Gate to Dam A Gate	0.003121	0.042525	0.045646	0.04657
Lacnor	Dumbell Lake gate to Dam A	0.001476	0.060696	0.062172	0.062172
Milliken	Tailing Management Area (Sheriff Creek Sanctuary)	0.001902	0.028084	0.029986	0.029986
	Gate 1 to Gate 2	0.002274	0.032206	0.034480	
Stanleigh	Tailing Management Area	0.001935	0.071417	0.073353	0.053916
Quirke	Gate to gate	0.006848	0.041048	0.047896	0.047896
	Gate 1 to peninsula	0.004803	0.066627	0.071430	
Panel	Tailing Management Area	0.003384	0.052695	0.056079	0.063754
Nordic	Gate to past Treatment Plant	0.001451	0.063308	0.064759	0.064759
Buckles		0.001883	0.046327	0.048210	0.048210
	Gate to Treatment Plant	0.002134	0.040689	0.042822	
Pronto	Tailing Management Area	0.003638	0.03363	0.037268	0.04005
Spanish American	Tailing Management Area	0.00169	0.036063	0.037753	0.037753
Esten Lake	Esten Boat Launch trail	0.001866	0.024304	0.026170	0.026170

^{*}Based on a radiation exposure period of 200 hours.

The casual access dose from measured radon and direct gamma radiation includes a background component. In order to calculate an incremental dose from walking near the TMAs, the dose as measured at Esten Lake must be subtracted from the dose measured at each TMA.

4.3 Radionuclides in Drinking Water and Fish

Measurements are available for some radionuclides in Elliot Lake drinking water (City of Elliot Lake, 2014, 2015), and in flesh samples from sport fish collected in lakes downstream



of the TMAs (Elliot Lake, Quirke Lake and McCarthy Lake) and in reference lakes (Minnow, 2005). Additional samples of treated drinking water were obtained by DES in August and November of 2016 and analyzed by SRC Environmental Analytical Laboratories for U-238 series radionuclides.

Based on the City of Elliot Lake measurements of uranium in treated drinking water (0.149 and 0.172 μ g/L), and the two DES measurements (both <0.1 μ g/L), the average uranium concentration in the treated water was 0.13 μ g/L, or approximately one thirteenth of the lake water concentrations used in previous dose assessments for Elliot Lake (EcoMetrix, 2011), where uranium was 1.7 μ g/L. Those concentrations included measured values for uranium and Pb-210, and detection limit values (<0.01 Bq/L) for Th-230, Ra-226 and Po-210. For the interim public dose assessment, the treated water concentration of uranium was 0.13 μ g/L and the concentrations of Th-234 and Th-230 were assumed to have the same ratios to U as reported for lake water by EcoMetrix (2011). Ra-226 in treated drinking water was estimated from the uranium concentration, based on the Ra/U ratio reported by Health Canada (2009) for the Elliot Lake water supply (0.015 Bq/L Ra per μ g/L U). Pb-210 and Po-210 were estimated from the Ra-226, based on the ratios for Elliot Lake water reported by EcoMetrix (2011). All the estimated radionuclide concentrations were below their limits of detection.

Health Canada (2009) has reported historical data for the Elliot Lake water supply (before treatment). The data show that concentrations of uranium and Ra-226 were relatively constant in 1995-96 when the record ends, at about 0.6 μ g/L and 0.007 Bq/L respectively. The concentrations for treated water, at present, are about one third of this level.

The treated drinking water concentrations include a background component. It is unclear what the background levels are in treated water. However, Health Canada (2009) reports that concentrations in Canadian water supplies range from <0.1 to 1 μ g/L for uranium, and from <0.005 to 0.02 Bq/L for Ra-226. In order to calculate an incremental dose from treated drinking water at Elliot Lake, a background uranium concentration of 0.1 μ g/L was assumed, and background concentrations of other radionuclides were estimated using ratios as described above. This implies a background concentration of 0.0015 Bq/L for Ra-226, which is unlikely to be detectable. Incremental dose can be calculated by subtracting the dose based on background concentrations from the dose based on Elliot Lake concentrations.

Using this low level of background is conservative, resulting in calculation of a small incremental exposure. Health Canada (2009) has suggested that the measured levels of radionuclides in the Elliot Lake water supply likely represent natural background rather than effects from uranium mining operations.

Average measured concentrations of uranium and Ra-226 in sport fish collected in 2004 from Elliot Lake (Minnow, 2005) were used in the interim public dose assessment. These concentrations, on a fresh weight basis, were, respectively, 0.0132 mg/kg and 0.20 Bq/kg for Elliot Lake fish, 0.0144 mg/kg and 0.238 Bq/kg for Quirke Lake fish, and 0.0148 mg/kg



and 0.32 Bq/kg for McCarthy Lake fish. Concentrations of Th-230 were estimated from uranium, and concentrations of Pb-210 and Po-210 were estimated from Ra-226, using the isotope ratios that were previously found in forage fish (EcoMetrix, 2011).

The survey of Elliot Lake residents by Rio Algom Limited in 2016 indicates that Elliot Lake, Quirke Lake and McCarthy Lake are the lakes most used for local fish consumption. The use proportions for these lakes from the survey were adjusted up to account for the people who did not know the lake fished, and to include the small fraction of people who used May Lake or Nordic Lake (collectively only 4% of users who knew the lake fished). The resulting proportions (50.4% Elliot, 28.3% Quirke, and 21.2% McCarthy) were used to weight the fish flesh concentrations across lakes, making a set of average concentrations for fish taken from exposed lakes, i.e. those downstream of TMAs.

The survey information also provided an estimate of the number of meals per year of fish from lakes downstream of TMAs (Elliot, Quirke, Nordic, McCabe, May and McCarthy), and this was converted to an intake rate for the representative person. The survey indicated an average of 7 meals per year for the typical Elliot Lake resident. For the interim dose estimate, using a meal size of 0.227 kg (fresh weight) (OMOECC, 2015), the intake rate of local fish was estimated at 1.59 kg/a.

The sport fish concentrations include a background component. Background levels were taken from sport fish collected in Dunlop Lake in 2004 (Minnow, 2005). For uranium and Ra-226, these levels were 0.01 mg/kg and 0.1 Bq/kg, respectively, on a fresh weight basis. Background concentrations for other radionuclides were estimated as described above. Incremental dose can be calculated by subtracting the dose based on background concentrations from the dose based on exposed lake concentrations in fish flesh.

4.4 Interim Public Dose Estimate

The interim public dose estimate for a representative person (adult) at Elliot Lake was calculated using radon and direct gamma measurements near TMAs, and radionuclide concentrations in treated drinking water and in sport fish flesh, as described above in Sections 4.2 and 4.3.

The casual access dose was calculated assuming 110.76 hours per year spent walking near the TMAs. The adult water intake of 1.5 L/d (Health Canada, 1995) was assumed to occur 365 days per year. This intake rate was applied to treated Elliot Lake drinking water. The adult intake of sport fish flesh from affected lakes was assumed to be 1.59 kg/year on a fresh weight basis.

Using these access and ingestion rates, the dose to human receptors was calculated as follows:

$$D_h = D_{r+q} + (C_w \cdot I_w + C_f \cdot I_f) \cdot DCF_i$$



where: D_h = human radiation dose (Sv/a)

 D_{r+g} = dose from radon and gamma, with TMA-specific values weighted by

proportion of local walking time spent at each TMA (Sv/a)

 C_w = activity concentration in drinking water (Bq/L)

 I_w = drinking water intake rate (L/a)

C_f = concentration in sport fish flesh, with lake-specific values weighted by

proportion of local intake arising from each lake (Bq/kg fw)

If = intake of sport fish flesh from affected lakes (kg fw / a)

 DCF_i = ingestion dose coefficient (Sv/Bq)

Ingestion dose coefficients were taken from ICRP Publication 72 (ICRP, 1996). The values provided by ICRP include dose contributions from progeny that may grow in over a lifetime following radionuclide ingestion. In addition, the values listed for parents and short-lived progeny have been combined to account for environmental ingrowth of progeny.

The dose limit for people (members of the public) is 1 mSv/a, as recommended in ICRP Publication 60 (ICRP, 1991). This is an incremental dose. Background radiation exposure, including natural and anthropogenic sources, is typically about 2 mSv/a (Health Canada, 2014). In addition, Health Canada (2014) has defined a dose constraint of 0.3 mSv/a as an incremental value above which dose management may be needed for naturally occurring radioactive materials. This is a conservative value which allows for exposure from other sources while still ensuring that incremental dose to a member of the public does not exceed the public dose limit.

The human doses calculated from measured radon, direct gamma, and radionuclide concentrations in affected areas include a natural background component. Therefore, the background component must be removed before comparison to the public dose limit, or to a dose constraint. The background component was estimated as described above, but using background values for radon, direct gamma and radionuclide concentrations, as described in **Sections 4.2** and **4.3**.

The interim total dose estimate, including background, as outlined in **Table 4-2**, is 0.030 mSv/a. The interim background dose estimate, as outlined in **Table 4-3**, is 0.018 mSv/a, and the incremental dose is 0.012 mSv/a. This is well below the public dose limit of 1 mSv/a, and also well below the dose constraint of 0.3 mSv/a.



Table 4.2: Estimation of Background-Inclusive Dose for a Representative Adult in Elliot Lake

Parameter	Units	U238/234	Th234+	Th230	Ra226	Rn222+	Pb210+	Po210	TOTAL
Water concentration	Bq/L	0.0032*	0.0026	0.0008	0.0020	0.0020	0.0059	0.0020	
Sport fish tissue conc.	Bq/kg (fw)	0.342#	0.169	0.099	0.237	0.024	0.430	0.624	
Ingestion rate water	L/a	547.5	547.5	547.5	547.5	547.5	547.5	547.5	
Ingestion rate fish	kg/a	1.59	1.59	1.59	1.59	1.59	1.59	1.59	
Exposure via water	Bq/a	1.75	1.41	0.42	1.07	1.07	3.20	1.07	
Exposure via fish	Bq/a	0.54	0.27	0.16	0.38	0.04	0.68	0.99	
Ingestion DCF adult	Sv/Bq	4.70E-08	3.40E-09	2.10E-07	2.80E-07	2.50E-10	6.91E-07	1.20E-06	
Dose via water	mSv/a	8.23E-05	4.80E-06	8.79E-05	2.99E-04	2.67E-07	2.21E-03	1.28E-03	3.97E-03
Dose via fish	mSv/a	2.56E-05	9.11E-07	3.31E-05	1.05E-04	9.41E-09	4.73E-04	1.19E-03	1.83E-03
Total ingestion dose	mSv/a	1.08E-04	5.72E-06	1.21E-04	4.04E-04	2.76E-07	2.69E-03	2.47E-03	5.80E-03
Casual access dose	mSv/a								2.39E-02
Total dose	mSv/a								2.97E-02

Table 4.3: Estimation of Background Dose for a Representative Adult

Parameter	Units	U238/234	Th234+	Th230	Ra226	Rn222+	Pb210+	Po210	TOTAL
Water concentration	Bq/L	0.0025*	0.0020	0.0006	0.0015	0.0015	0.0045	0.0015	
Sport fish tissue conc.	Bq/kg (fw)	0.246#	0.121	0.071	0.100	0.010	0.182	0.264	
Ingestion rate water	L/a	547.5	547.5	547.5	547.5	547.5	547.5	547.5	
Ingestion rate fish	kg/a	1.59	1.59	1.59	1.59	1.59	1.59	1.59	
Exposure via water	Bq/a	1.35	1.09	0.32	0.82	0.82	2.46	0.82	
Exposure via fish	Bq/a	0.39	0.19	0.11	0.16	0.02	0.29	0.42	
Ingestion DCF adult	Sv/Bq	4.70E-08	3.40E-09	2.10E-07	2.80E-07	2.50E-10	6.91E-07	1.20E-06	
Dose via water	mSv/a	6.33E-05	3.70E-06	6.76E-05	2.30E-04	2.05E-07	1.70E-03	9.86E-04	3.05E-03
Dose via fish	mSv/a	1.84E-05	6.56E-07	2.38E-05	4.45E-05	3.98E-09	2.00E-04	5.03E-04	7.90E-04
Total ingestion dose	mSv/a	8.17E-05	4.35E-06	9.15E-05	2.74E-04	2.09E-07	1.90E-03	1.49E-03	3.84E-03
Casual access dose	mSv/a								1.45E-02
Total dose	mSv/a								1.83E-02

⁺ indicates that progeny contributions are included in the DCF

⁺ indicates that progeny contributions are included in the DCF



5.0 CONCLUSIONS AND RECOMMENDATIONS

Based on the interim public dose calculations, it may be concluded that:

- Public dose to the representative person is approximately 0.012 mSv/a, after correction for background exposure.
- This value is based on available measurements of radon and direct gamma near TMAs, and U-238 series radionuclides in treated drinking water and sport fish, as well as critical group survey information and several assumptions for exposure factors.

The public dose estimation may be refined in the future based on information from critical group surveys and from the monitoring program.

Preliminary recommendations for the monitoring program to support future public dose estimates include:

- Prior to the next reporting cycle, update the sampling and analysis of U-238 series
 radionuclides in sport fish collected from lakes of the Serpent River watershed most
 used by sport fishers (Elliot, Quirke and McCarthy lakes); target lake trout and
 walleye, which were the species most consumed according to the survey of Elliot
 Lake residents. Smallmouth bass and northern pike were next in order of
 preference, and may also be used.
- In subsequent cycles, consider whether the resident survey or components of the monitoring program may need to be updated, based on possible demographic changes in the community, changes in waste management operations, or trends observed in the watershed monitoring program.

The information from the resident survey and monitoring programs is expected to be used in public dose estimation as described in Section 4.4. Public dose estimates may be revised in the future as the relevant information is updated.



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Appendix A Questionnaire and Results for Survey of Elliot Lake Residents



q5.1t recode. How many years have you live	ved in the	area arour	nd Elliot La	ke?							
Sample Size	300										
Less than 1											
Column %	1										
1 - 5											
Column %	8										
6 - 10											
Column %	9										
11 - 15											
Column %	9										
16 - 20											
Column %	14										
21 - 25											
Column %	8										
26 - 30											
Column %	6										
31 - 35											
Column %	6										
36+											Ì
Column %	37										
DK/NA											
Column %	1										
q5.1t_mean. How many years have you live	d in the a	rea around	l Elliot Lak	e?							
<u> </u>											
Sample Size	296										
Mean	29										
Std. Dev.	18										
Q5.2. Do you ever eat fish that was caught	in local lal	es - in oth	er words.	from eithe	r Elliot La	ke, McCarth	v Lake, M	av Lake. N	ordic Lake	or Quirke	e Lake?
, , , , , , , , , , , , , , , , , , , ,			ĺ					,			
Sample Size	300	Meals/Yea	Г	Weighted /	Average Me	eals per Year					
Yes - at least three times a week or more											
Column %	1	156		7							
Yes - around once a week on average over the	vear										
Column %	8	52									
Yes - about once a month on average over the	vear										
Column %	14	12									
Yes - a few times a year											
Column %	41										
No. never		3									
		3									
Column % DK/NA	31	0									
Column %											
Column % DK/NA	31	0									
Column % DK/NA Column %	31	0	of the loca	I fish that	you eat co	mes from?					
Column % DK/NA	31	0	of the loca	I fish that	you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where	31 5 would you	0 0 say most o		I fish that	you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where	31 5 would you	0		I fish that	you eat co	omes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where sample Size Elliot Lake	31 5 would you	0 0 say most o	that know		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where sample Size Elliot Lake Column %	31 5 would you	0 0 say most o	that know		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where sample Size Elliot Lake Column % McCarthy Lake	31 5 would you 192 30	0 0 say most o % of those 48.3	that know 50.4		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where we sample Size Elliot Lake Column % McCarthy Lake Column %	31 5 would you	0 0 say most o	that know 50.4		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where the sample Size Elliot Lake Column % McCarthy Lake Column % McCabe Lake	31 5 would you 192 30	0 0 say most 0 % of those 48.3 20.3	50.4 21.2		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where we sample Size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % McCabe Lake Column %	31 5 would you 192 30	0 0 say most o % of those 48.3	50.4 21.2		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where some size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % May Lake	31 5 would you 192 30 13	0 say most of % of those 48.3 20.3	50.4 21.2		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where with the lakes I just mentioned, where with the lake I just mentioned, where with the lake I just mentioned, where with lake I just mentioned, where with lake I just mentioned, where with lake I just mentioned in the lake I just mentioned	31 5 would you 192 30	0 say most of % of those 48.3 20.3	50.4 21.2		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where sample Size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % McColumn % McCabe Lake Column % My Lake Column % Nordic Lake	31 5 would you 192 30 13	0 say most of % of those 48.3 20.3	that know 50.4 21.2 omit		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where some size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % May Lake Column % Nordic Lake Column % Nordic Lake Column %	31 5 would you 192 30 13	0 say most of % of those 48.3 20.3	that know 50.4 21.2 omit		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where some size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % May Lake Column % May Lake Column % May Lake Column % Mordic Lake Column % Ourick Lake Column % Quirke Lake	31 5 would you 192 30 13 0	0 say most c % of those 48.3 20.3 0.0 3.4	that know 50.4 21.2 omit		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where some sample Size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % May Lake Column % May Lake Column % Mordic Lake Column % Quirke Lake Column % Quirke Lake Column %	31 5 would you 192 30 13	0 say most of % of those 48.3 20.3	that know 50.4 21.2 omit		you eat co	mes from?					
Column % DK/NA Column % Q5.3. Of the lakes I just mentioned, where some size Elliot Lake Column % McCarthy Lake Column % McCabe Lake Column % May Lake Column % May Lake Column % May Lake Column % Mordic Lake Column % Ourick Lake Column % Quirke Lake	31 5 would you 192 30 13 0	0 say most of % of those 48.3 20.3 0.0 3.4 0.8	that know 50.4 21.2 omit		you eat co	mes from?					



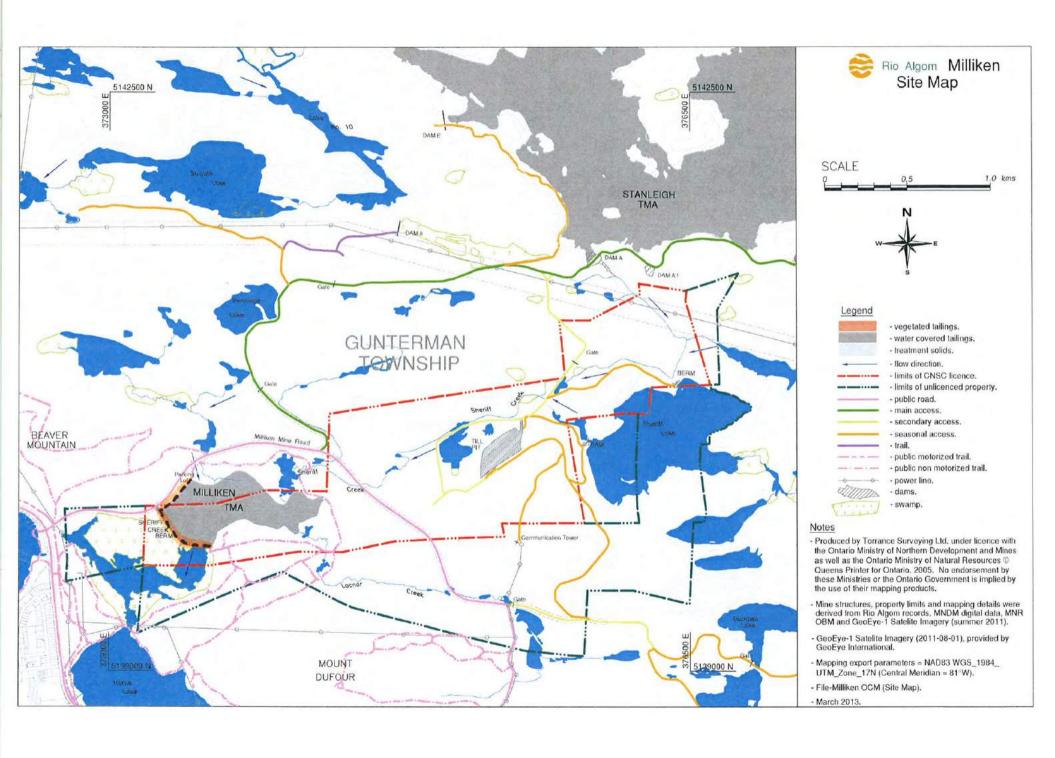
Q5.4. What species of fish caugl	nt from local lakes would yo	u say you eat mo	ost often?				
Sample Size	192						
Lake trout							
Column %	43						
Brook/Speckle trout							
Column %	2						
Rainbow trout							
Column %	4						
Northern pike							
Column %	6						
Smallmouth bass							
Column %	7						
Walleye/Pickerel							
Column %	21						
Splake							
Column %	1						
Perch							
Column %	3						
Whitefish							
Column %	1						
Sturgeon							
Column %	1						
Other							
Column %	13						
Column 78	13						
Q5.5. Do you have any children	under the age of 16 in your	household that a	at fich that was	caught in local	lakas in oth	or words	
from either Elliot Lake, McCart					iakes - iii oui	er words,	
nom emier Emot Lake, wccan	illy Lake, wiccabe Lake, way	Lake, Noruic La	ike of Quilke La	ine :			
Sample Size	300						
Yes	300						
Yes Column %	7						
	/						
No or or	00						
Column %	92						
DK/NA							
Column %	2						

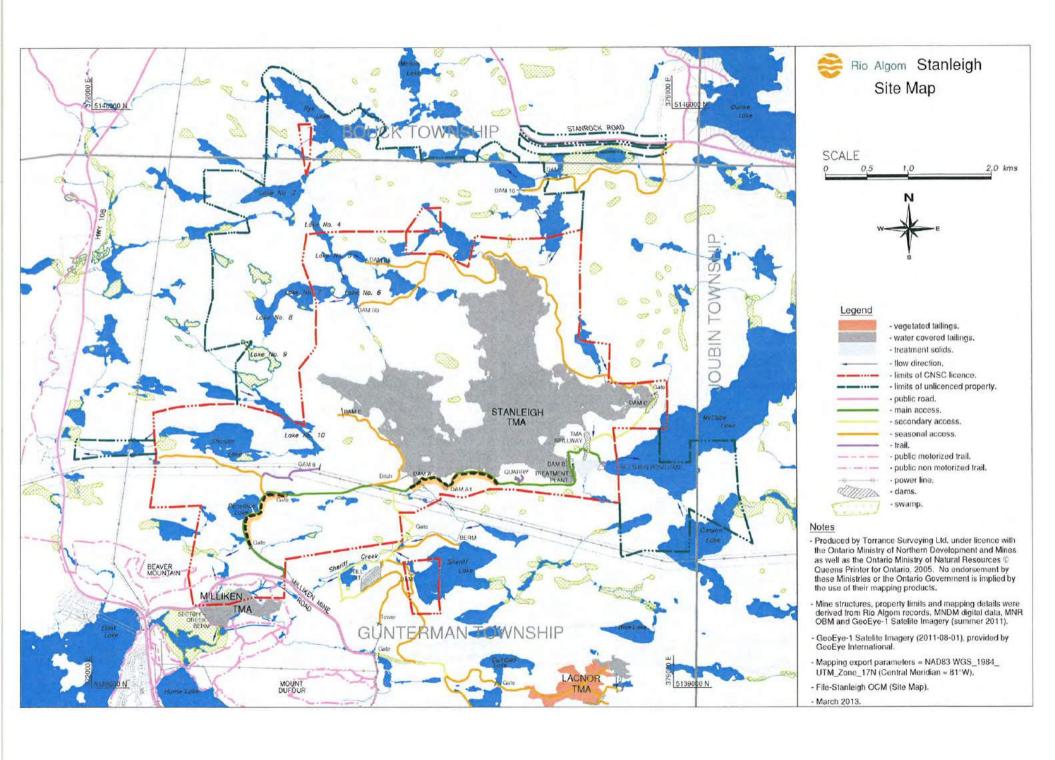


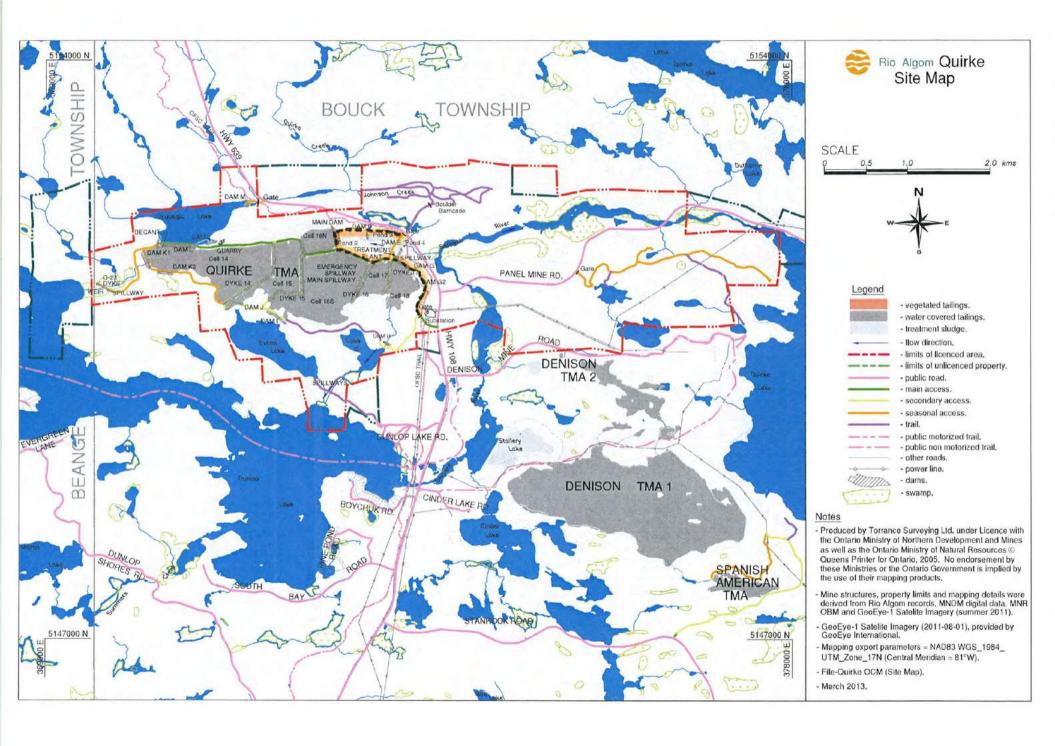
								1			
									0.4.4		<u> </u>
q5.7 recode. How many hours per we											include
Quirke, Panel, Spanish American, S	tanieign, Millikei	n and Sne	гіт сгеек	Park, Laci	nor, Noraid	, Buckles, I	ronto, De	nison and	Stanrock)		
Sample Size	300										
Zero	300					l .					
Column %	59					l I					
1 - 5	33					l					
Column %	18										-
6 - 10	10										
	7										
Column %	7					I					
11 - 15											-
Column %	3										
16 - 20											
Column %	1										
21+											
Column %	1										
DK/NA											
Column %	12										
q5.7_mean. How many hours per wee	ek would you say	you spen	d walking	or hiking	around the	e closed mi	ne propert	ies in the	area?		
Sample Size		Hours per									
Mean	2.13	110.76									
Std. Dev.	5.39										
Q5.8. Considering the mine propertie	s I just mentioned	d, on whic	ch one wo	uld you sa	y you wall	or hike the	most?				
Sample Size	124	% of those	that know								
Quirke											
Column %	10	12.6									i i
Panel	10	12.0									
Column %	6	8.4									
Spanish American		0.7									
Column %	0	0.0				l I					
Stanleigh	0	0.0									_
Column %	7	9.5									
	- /	9.0				l I					
Milliken / Sheriff Creek Park	25	45.0									
Column %	35	45.3				<u> </u>					
Lacnor											-
Column %	0	0.0									
Nordic											
Column %	7	9.5									
Buckles											
Column %	0	0.0									
Pronto										<u> </u>	
Column %	0	0.0									
Denison											
Column %	6	7.4									
Stanrock											
Column %	6	7.4									
DK/NA											
Column %	23										Î
Q5.9. Do you have any children unde	rage 16 who spe	end time w	valking or	hiking aro	und the cl	osed mine	properties	in the are	a?		
	go .o wno spe	a anno v			alo 61	2304 //////	oportioo				1
Sample Size	300										
Yes	300										
				I		I					-
Column %	6										
No .											
Column %	92										
DK/NA											
Column %	2										

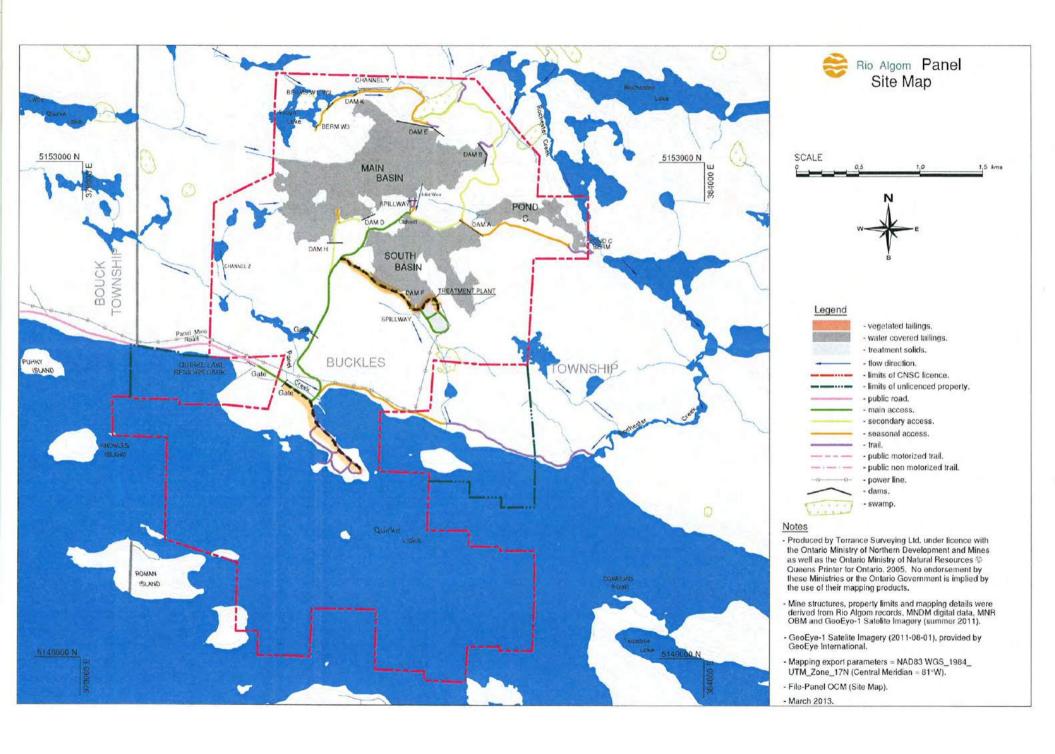


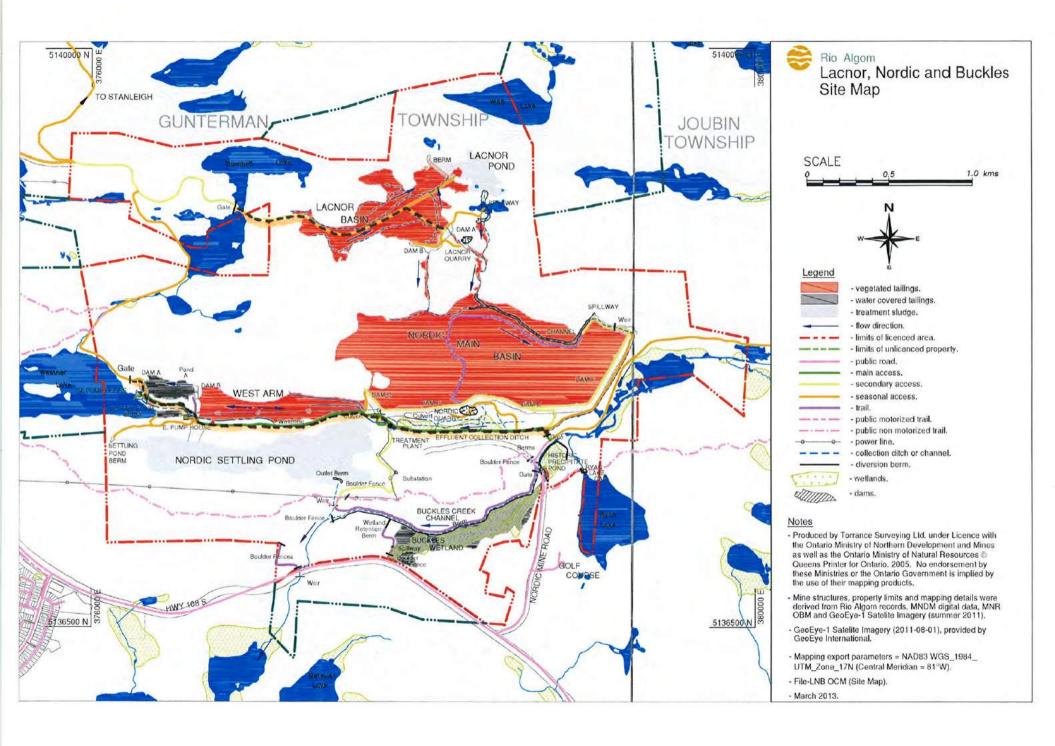
Appendix B Site Maps with Walking Surveys for Radon and Gamma

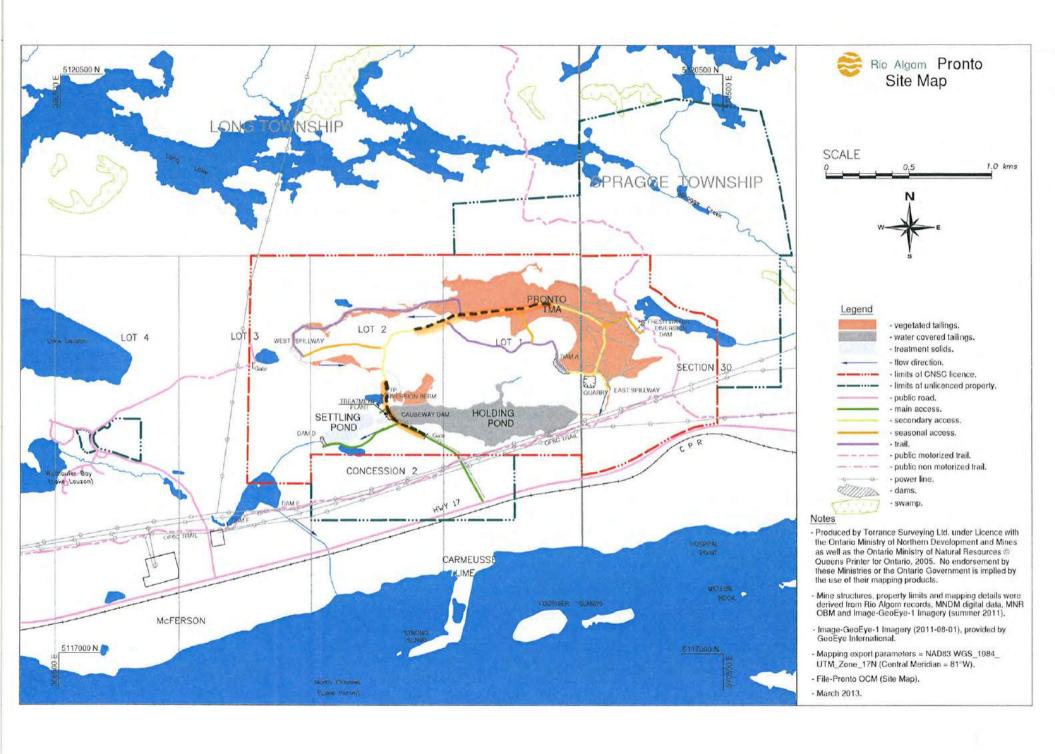


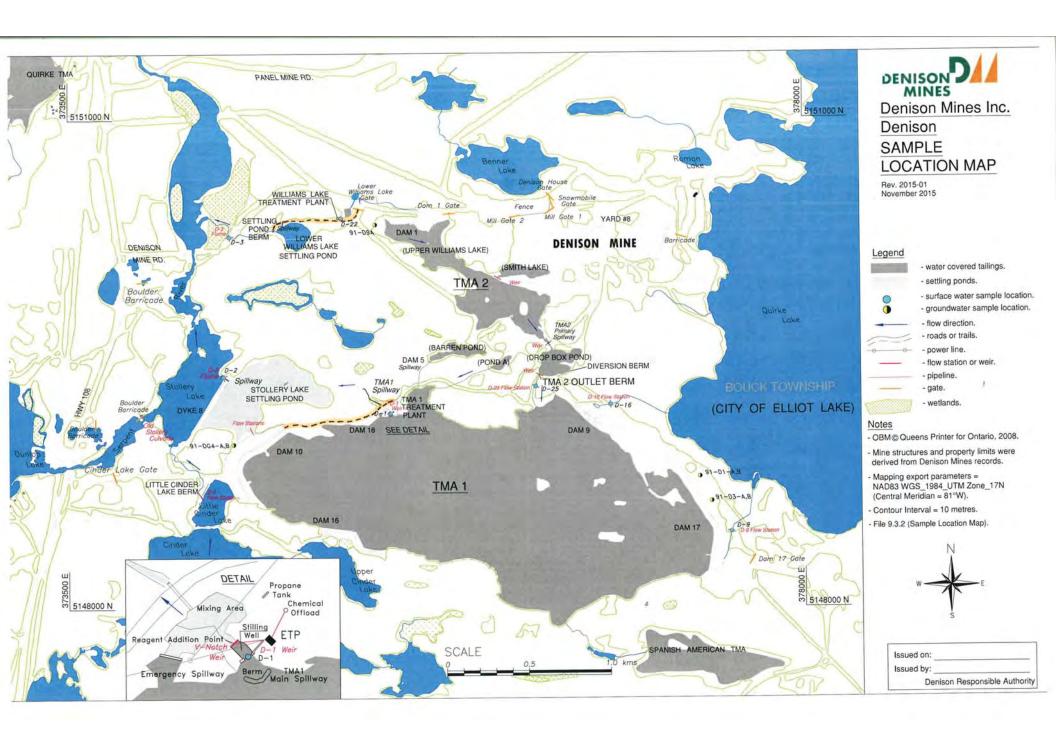


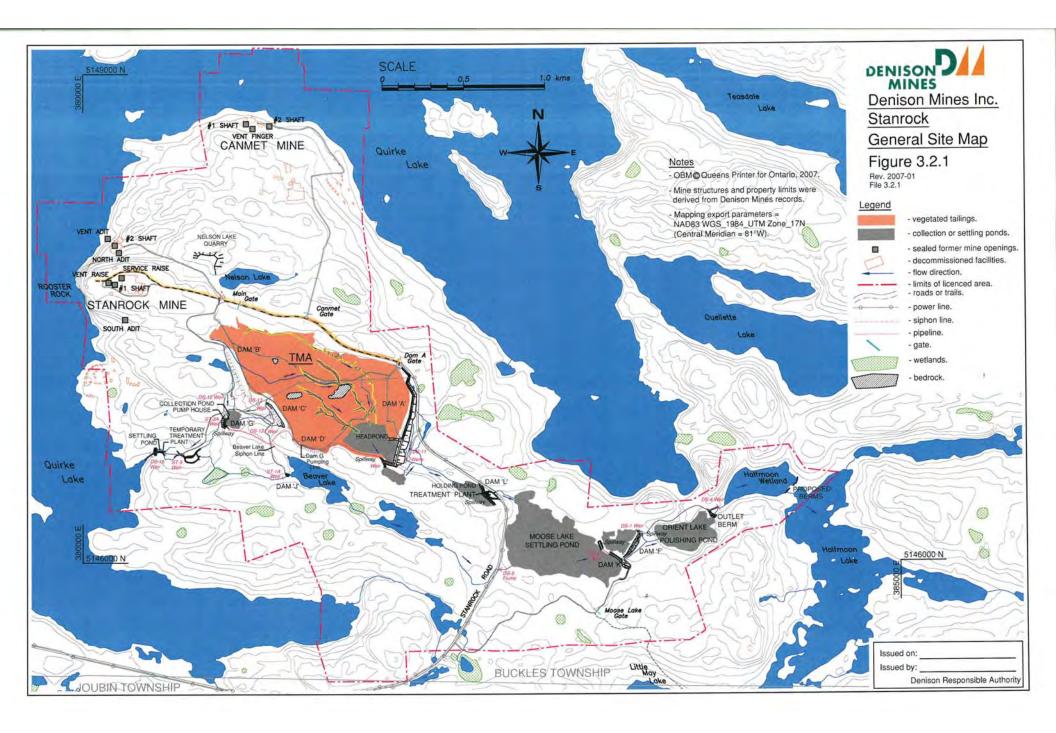














Appendix C Certificates of Analysis for Drinking Water



102 - 422 Downey Road, Saskatoon, SK Canada S7N 4N1

T: 306-933-6932 **F**: 306-933-7922 **Toll-free**: 1-800-240-8808 **E**: analytical@src.sk.ca

www.src.sk.ca/analytical

SRC Group # 2016-10582

Sep 21, 2016

Denison Environmental Services 1 Horne Walk, Suite 200 Elliot Lake, ON P5A 2A5 Attn: Valerie Kilp

Date Samples Received: Sep-06-2016 Client P.O.: 107732

This is a final report.

Lab Section 1 results have been authorized by Keith Gipman, Supervisor

Lab Section 2 results have been authorized by Melissa Tackaberry-Syed, Supervisor

Lab Section 3 results have been authorized by Pat Moser, Supervisor

Lab Sections 4 and 5 results have been authorized by Vicky Snook, Supervisor

Lab Section 6 results have been authorized by Marion McConnell, Supervisor

- * Routine methods follow recognized procedures from sources such as
 - * Standard Methods for the Examination of Water and Wastewater APHA AWWA WEF
 - * Environment Canada
 - * US EPA
 - * CANMET

- * Samples will be kept for 30 days after the final report is sent. Please contact the lab if you have any special requirements.
- * Additional information is available upon request.

^{*} Test methods and data are validated by the laboratory's Quality Assurance Program.

^{*} The results reported relate only to the test samples as provided by the client.



102 - 422 Downey Road, Saskatoon, SK Canada S7N 4N1

T: 306-933-6932 F: 306-933-7922 Toll-free: 1-800-240-8808 E: analytical@src.sk.ca

www.src.sk.ca/analytical

SRC Group # 2016-10582 Sep 21, 2016

Denison Environmental Services

1 Horne Walk, Suite 200 Elliot Lake, ON P5A 2A5 Attn: Valerie Kilp

Date Samples Received: Sep-06-2016 Client P.O.: 107732

27799 08/31/2016 DWW *WATER*

Analyte	Units	27799	
Lab Section 2 (ICP)			
Uranium	ug/L	<0.1	
Lab Section 4 (Radiocher	mistry)		
Lead-210	Bq/L	<0.02	
Polonium-210	Bq/L	<0.005	
Radium-226	Bq/L	<0.005	
Thorium-230	Bq/L	<0.01	

Symbol of "<" means "less than". This indicates that it was not detected at level stated above.



102 - 422 Downey Road, Saskatoon, SK Canada S7N 4N1

T: 306-933-6932 F: 306-933-7922 Toll-free: 1-800-240-8808 E: analytical@src.sk.ca

www.src.sk.ca/analytical

SRC Group # 2016-14713

Dec 14, 2016

Denison Environmental Services 1 Horne Walk, Suite 200 Elliot Lake, ON P5A 2A5 Attn: Valerie Kilp

Date Samples Received: Dec-01-2016 Client P.O.: 107732

This is a final report.

Lab Section 1 results have been authorized by Keith Gipman QP, Supervisor

Lab Section 2 results have been authorized by Melissa Tackaberry-Syed QP, Supervisor

Lab Section 3 results have been authorized by Pat Moser QP, Supervisor

Lab Sections 4 and 5 results have been authorized by Vicky Snook QP, Supervisor

Lab Section 6 results have been authorized by Marion McConnell QP, Supervisor

QP: Qualified Person in accordance with the Saskatchewan Environmental Code, Corrective Action Plan Chapter, for the purposes of certifying a laboratory analysis

- * Test methods and data are validated by the laboratory's Quality Assurance Program.
- * Routine methods follow recognized procedures from sources such as
 - * Standard Methods for the Examination of Water and Wastewater APHA AWWA WEF
 - * Environment Canada
 - * US EPA
 - * CANMET
- * The results reported relate only to the test samples as provided by the client.
- * Samples will be kept for 30 days after the final report is sent. Please contact the lab if you have any special requirements.
- * Additional information is available upon request.



102 - 422 Downey Road, Saskatoon, SK Canada S7N 4N1

T: 306-933-6932 F: 306-933-7922 Toll-free: 1-800-240-8808 E: analytical@src.sk.ca

www.src.sk.ca/analytical

SRC Group # 2016-14713

Dec 14, 2016

Denison Environmental Services

1 Horne Walk, Suite 200 Elliot Lake, ON P5A 2A5

Attn: Valerie Kilp

Date Samples Received: Dec-01-2016 Client P.O.: 107732

40387 11/29/2016 DWW *WATER*

Analyte	Units	40387	
Lab Section 2 (ICP)			
Uranium	ug/L	<0.1	
Lab Section 4 (Radioche	mistry)		
Lead-210	Bq/L	<0.02	
Polonium-210	Bq/L	<0.005	
Radium-226	Bq/L	<0.005	
Thorium-230	Bq/L	<0.01	

Symbol of "<" means "less than". This indicates that it was not detected at level stated above.